

Potential impacts of future land use and climate change on the Red List status of the Proteaceae in the Cape Floristic Region, South Africa

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Abstract

Using spatial predictions of future threats to biodiversity, we assessed for the first time the relative potential impacts of future land use and climate change on the threat status of plant species. We thus estimated how many taxa could be affected by future threats that are usually not included in current IUCN Red List assessments. Here, we computed the Red List status including future threats of 227 Proteaceae taxa endemic to the Cape Floristic Region, South Africa, and compared this with their Red List status excluding future threats. We developed eight different land use and climate change scenarios for the year 2020, providing a range of best- to worst-case scenarios. Four scenarios include only the effects of future land use change, while the other four also include the impacts of projected anthropogenic climate change (HadCM2 IS92a GGa), using niche-based models. Up to a third of the 227 Proteaceae taxa are uplisted (become more threatened) by up to three threat categories if future threats as predicted for 2020 are included, and the proportion of threatened Proteaceae taxa rises on average by 9% (range 2–16%), depending on the scenario. With increasing severity of the scenarios, the proportion of Critically Endangered taxa increases from about 1% to 7% and almost 2% of the 227 Proteaceae taxa become Extinct because of climate change. Overall, climate change has the most severe effects on the Proteaceae, but land use change also severely affects some taxa. Most of the threatened taxa occur in low-lying coastal areas, but the proportion of threatened taxa changes considerably in inland mountain areas if future threats are included. Our approach gives important insights into how, where and when future threats could affect species persistence and can in a sense be seen as a test of the value of planned interventions for conservation.

Key words: biodiversity conservation, Cape Floristic Region, climate change, extinction risk, IUCN Red List, land use change, niche-based models, Proteaceae, threatened species

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Introduction

Land use and climate change are considered to be the major drivers of global biodiversity change in terrestrial ecosystems in the near future (Sala *et al.*, 2000). Habitat

destruction and degradation, mostly because of agriculture, invasive alien species and urbanization are the main past and present causes of species endangerment and extinction (Czech & Krausman, 1997; Wilcove *et al.*, 1998). However, anthropogenic climate change increasingly imposes a major additional threat, causing species range shifts and losses, population declines and extinctions (Hughes, 2000; McCarty, 2001; Parmesan & Yohe, 2003; Root *et al.*, 2003).

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Similarly, in the Cape Floristic Region (CFR), a global biodiversity hotspot comprising one of the world's six floristic kingdoms (Cowling *et al.*, 1997; Myers *et al.*, 2000), habitat transformation through agriculture, invasive alien plants and urbanization has been identified as a major past, present and future threat to biodiversity (Rebelo, 1992; Richardson *et al.*, 1996; Rouget *et al.*, 2003c; Latimer *et al.*, 2004). In addition, anthropogenic climate change is likely to be a major future threat to the biodiversity in the CFR (Rutherford *et al.*, 1999; Midgley *et al.*, 2002, 2003).

The threat status of the world's species is typically assessed based on what we currently know about species distributions, population sizes and population declines. The IUCN has produced standardized Red List criteria that are the most commonly used tools to identify and classify species of conservation concern (Lamoreux *et al.*, 2003). Although the criteria allow for future threats to be included in IUCN Red List assessments (IUCN, 2001, 2003), lack of data or uncertainty about future threats means that most assessments are based on current species distributions and population sizes or past population declines. The impact of future threats, particularly threats such as climate change, on the threat status of most species remains largely unknown (Burgman, 2002). It is, therefore, difficult to determine whether the inclusion of future threats would result in a substantial change in the number and Red List status of threatened species, the types of species that are threatened or the distribution of these species, with concomitant implications for conservation planning and action.

This study assesses how the threat status of plant species could change if environmental conditions predicted by land use and climate change models for the year 2020 are incorporated into a Red List assessment. Our aim is not to test the IUCN Red List Categories and Criteria, but to estimate how many species are likely to be affected by future land use and climate change, threats that are usually not included in current IUCN Red List assessments. Such a study provides a first estimation of the relative importance of future threats for species conservation efforts.

The impact of future land use change and, recently, climate change has been translated into crude estimates of extinction risk elsewhere (Gaston *et al.*, 2003; Thomas *et al.*, 2004). However, to the best of our knowledge, nobody has yet calculated a Red List status for a set of species, taking into account the combined impacts of land use and climate change in the near future, and compared it with their current Red List status. In the CFR, high-resolution species distribution data from the Protea Atlas Project (Rebelo, 2001), spatial predictions of land use change (Rouget *et al.*, 2003c) and spatial,

species-specific predictions of climate change (Midgley *et al.*, 2003) provide the opportunity to assess the potential impacts of future land use and climate change on the Red List status of the Proteaceae, a characteristic plant family of the predominant fynbos biome in the CFR.

Here, we compute the Red List status of 227 Proteaceae taxa endemic to the CFR, for which species-specific climate change predictions were available for 2020, including future threats, and compare this with their Red List status excluding future threats. The timeframe of this study corresponds to that of the Cape Action for People and the Environment programme (<http://capeaction.org.za>) that seeks to conserve the biodiversity of the CFR (Cowling *et al.*, 2003). We developed eight different land use and climate change scenarios for the year 2020. We consider these to provide a reasonable range of best- to worst-case scenarios of environmental conditions for the near future in the CFR. Four scenarios include only the effects of habitat transformation and, indirectly, habitat conservation, ranging from best- to worst-case estimates of these parameters (White *et al.*, 1997). An additional four scenarios also include the impacts of climate change (Midgley *et al.*, 2003).

Specifically we ask: How does future land use and climate change impact on the number and Red List status of threatened Proteaceae in the CFR, what are their relative potential impacts and which taxa face the highest threat status or largest changes in their threat status? Furthermore, we assess whether or not land use and climate change impact differently on taxa from different regions and altitudes as has been shown elsewhere (Rebelo, 1992; Richardson *et al.*, 1996; Midgley *et al.*, 2002, 2003). Finally, we identify areas in which the number of threatened taxa increases because of future land use and climate change as projected here – information potentially useful for proactive conservation planning and action.

Materials and methods

Study area

The CFR covers an area of approximately 87 000 km² at the southern tip of Africa (Fig. 1) and contains some 9030 vascular plant species, of which nearly 70% are endemic (Goldblatt & Manning, 2002). Fynbos, the predominant vegetation type in the CFR, is an evergreen, fire-prone shrubland mainly characterized by three plant families: Proteaceae, Ericaceae and Restionaceae (Cowling *et al.*, 1997).

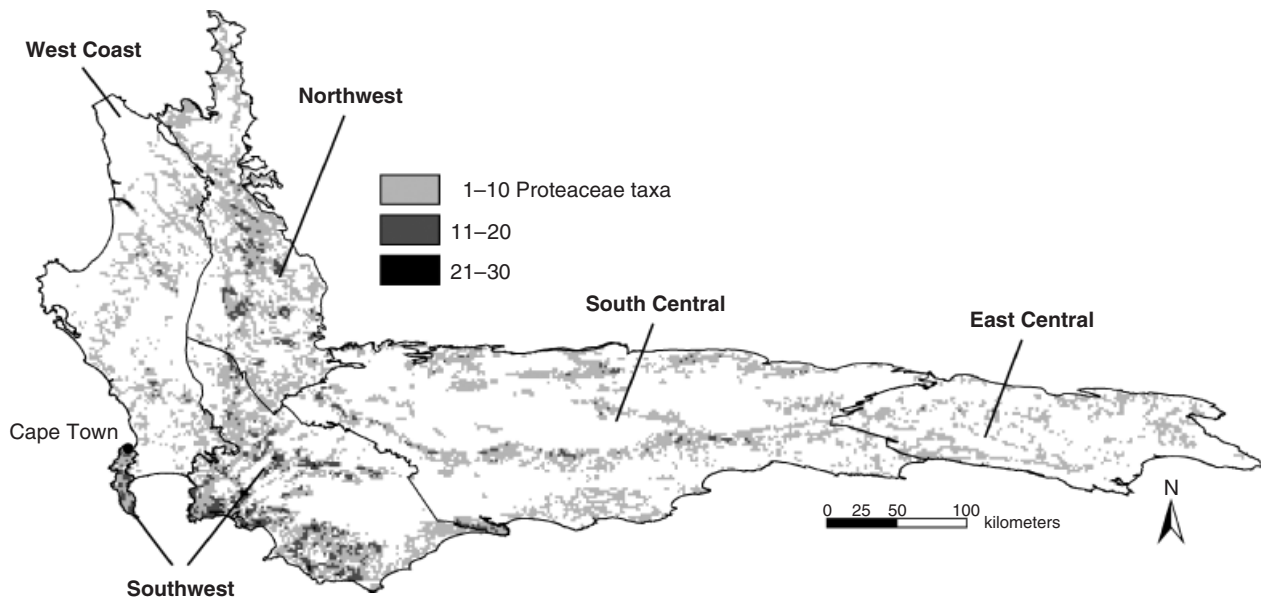


Fig. 1 South Africa's Cape Floristic Region showing phylogeographic regions and species richness summarized per 1 min grid cell (2.9 km²), as recorded by the Protea Atlas Project in October 2003, for the 227 Proteaceae taxa considered in this study.

Scenarios

Eight different land use and climate change scenarios were developed for 2020 (Fig. 2). Four scenarios, labelled –CC1 to –CC4, consider only the effects of habitat transformation and, indirectly, habitat conservation, ranging from best- to worst-case estimates of these parameters. Similar in their consideration of land use change, an additional four scenarios, labelled +CC1 to +CC4, include the impacts of climate change on species distributions based on a single climate change scenario (HadCM2 IS92a GGa). For the Red List assessment excluding and including future threats, information on species distributions and habitat conservation and transformation was used as follows.

Current and future species distributions

Current species distributions were obtained from the South African National Biodiversity Institute's (SANBI) Protea Atlas Project (<http://protea.worldonline.co.za>). This is one of the most comprehensive species distribution data sets in Africa with relatively low levels of sampling bias (Lombard *et al.*, 2003). It currently contains some 250 000 records for 377 species, subspecies and varieties (hereafter species or taxa) of Proteaceae (Rebello, 2001). For our Red List assessment at a regional level (Gärdenfors *et al.*, 2001), all 227 Proteaceae taxa endemic to the CFR for which there were species-specific climate change predictions avail-

able were selected, as recorded by the Protea Atlas Project in October 2003 (Fig. 1). They represent terminal taxa – species, subspecies and varieties – of 11 genera (nomenclature follows Rebello, 2001). All taxa were assigned to one of five phylogeographic regions (Fig. 1) (Goldblatt & Manning, 2002) and to one of four altitudinal zones based on where they are predominantly recorded at present.

Potential current and future species distributions for 2020 were obtained from SANBI's Climate Change Research Group (Midgley *et al.*, 2003). These were modelled at a 1 min scale using Generalized Additive Models (GAM), which determine the environmental limits of an entity with a given spatial distribution by matching its distribution to surfaces of environmental parameters (Yee & Mitchell, 1991). Five bioclimatic and three edaphic variables, all considered critical to plant survival (Midgley *et al.*, 2003), were used to build stepwise GAMs of environmental constraints for each species in S-Plus (MathSoft S-Plus Version 2000). Probability of occurrence from GAMs was mapped using ArcView software (ESRI ArcView Version 3.2). Then species-specific probability cut-off points were determined by maximizing jointly the percentage of presence and absence correctly predicted (Pearce & Ferrier, 2000; Thuiller *et al.*, 2003). Further details on bioclimatic modelling and its limitations are provided elsewhere (Pearson *et al.*, 2004; Thuiller, 2004; Thuiller *et al.*, 2004). Potential future species distributions for 2020 were derived using climate change projections

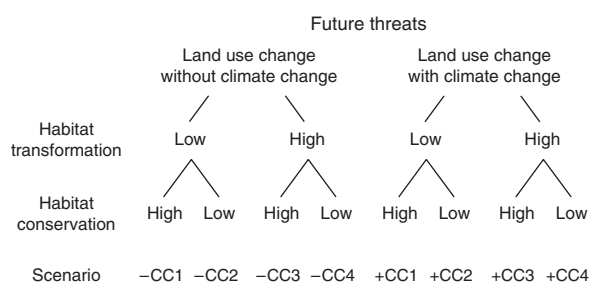


Fig. 2 Overview of the future scenarios and their labelling (scenarios increase in severity from left to right: -CC is without climate change; +CC is with climate change). The scenarios -CC1 to -CC4 and +CC1 to +CC4 are similar in their consideration of land use change, but the latter include the impacts of climate change on species distributions based on a single climate change scenario (HadCM2 IS92a GGa).

originating from the Global Circulation Model HadCM2 of the UK Hadley Centre for Climate Prediction and Research for southern Africa for 2050. The applied HadCM2 IS92a GGa scenario of the model predicts an average increase of 0.7°C ($\pm 0.1^{\circ}\text{C}$, range $0.5\text{--}1.0^{\circ}\text{C}$) in mean annual temperature and an average decrease of 41 mm ($\pm 25\text{ mm}$, range $7\text{--}295\text{ mm}$) in mean annual precipitation for the CFR for 2020 assuming a constant rate of climate change through to 2050.

Future species distributions for the Red List assessment including climate change impacts were finally determined by deleting those records from the original species distribution data set that are outside the modelled future ranges. This approach involves two simplifying assumptions, which may introduce some uncertainty. Firstly, all plants that currently occur outside the modelled future ranges will be extinct by 2020. As most Proteaceae are relatively long-lived species with life spans in excess of 15 years (Cowling *et al.*, 1997; Rebelo, 2001) this is unlikely. However, we assumed that viable populations would not persist permanently outside the potential future ranges, and we thus treated plants in areas where they are eventually committed to extinction as extinct by 2020. It is noteworthy that current species records may be outside the modelled future ranges because the affected areas become bioclimatically unsuitable or simply because the models are a poor fit to the data. We did, however, not attempt to correct for this. Secondly, we assumed that Proteaceae would not migrate into new areas, which will become bioclimatically suitable by 2020, because most of the Proteaceae are localized dispersers whose dispersal is closely linked to the fire cycle in the fynbos biome (Cowling *et al.*, 1997; Rebelo, 2001). It is unlikely that, without human facilitation, significant migration will occur across the CFRs fragmented landscapes by 2020. According to the

bioclimatic models for all the study taxa, 25.9% of the current species records are outside the modelled future ranges for 2020. While seven taxa do not lose any of their records, four taxa lose all of them.

Current and future habitat conservation and transformation

Information on current and future conservation areas (Fig. 3b, c) was used to identify areas that are unlikely to be affected by future habitat transformation. Our worst-case future conservation scenario assumes that the existing conservation area network remains unchanged until 2020. A Geographic Information Systems (GIS) layer of existing reserves was obtained from the Western Cape Nature Conservation Board's Conservation Planning Unit (<http://cpu.uwc.ac.za>). It includes statutory reserves such as national parks, provincial nature reserves and local authority reserves, as well as some nonstatutory reserves such as national heritage sites and private nature reserves (Rouget *et al.*, 2003b). Our best-case future conservation scenario assumes that this existing conservation area network will be expanded by the successful implementation of three proposed megareserves before 2020 (Cowling *et al.*, 1999; Younge, 2000). The proposed megareserves each cover between 400 000 and 600 000 ha of currently untransformed habitat, and extend across climatic, biotic and altitudinal gradients to alleviate future climate change impacts (Cowling *et al.*, 1999; Younge, 2000). Their purpose is to protect areas extensive and diverse enough to maintain top predators, megaherbivores, natural fire regimes and other landscape-scale processes (Cowling *et al.*, 1999; Younge, 2000). They will use a wide range of mechanisms including land use zoning, statutory and nonstatutory reserves to meet this purpose. For this study, we obtained GIS layers of the latest available, but still preliminary, outlines of the proposed Gouritz, Cederberg and Baviaanskloof megareserves from the Western Cape Nature Conservation Board's Conservation Planning Unit and Baviaanskloof Megareserve Project. The megareserves are expected to be implemented before 2020 (Younge, 2000), which corresponds to the timeframe of this study, and thus provide, together with the existing conservation areas, a first estimate for future habitat conservation in the CFR.

Information on current and future habitat transformation (Fig. 3a) was used as a surrogate for population reduction in the Red List assessment (IUCN, 2001, 2003). A GIS layer on current habitat transformation, defined as urbanized or cultivated areas (including forestry plantations) and medium- or high-density stands of invasive alien trees and shrubs (Rouget *et al.*, 2003c), was provided by the Protea Atlas Project.

It was originally derived from LANDSAT TM imagery captured in 1997 and 1998 (Lloyd *et al.*, 1999; Rouget *et al.*, 2003c).

Only currently untransformed habitat outside conservation areas can be transformed in the future because we assumed that management within all current and future conservation areas would prevent further habitat transformation. Best- and worst-case scenarios, considered to cover a realistic range of potential future land use change, were developed for future habitat transformation (Fig. 3a). Information on future habitat transformation was derived from 1 min grid layers (2.9 km² cell size) on future threats to biodiversity developed by Rouget *et al.* (2003c). They identified currently untransformed areas threatened by urbanization, using a simplistic rule-based modelling approach and currently untransformed areas threatened by agriculture and invasive alien trees and shrubs, using a statistical modelling approach. They classified all areas as follows: not threatened, moderately or highly threatened by urbanization; not threatened or threatened by agriculture and moderately or highly threatened by alien plant invasions (Rouget *et al.*, 2003c).

Not all areas threatened by future habitat transformation will be transformed by 2020. Thus, it was necessary to identify areas most likely to be transformed in 2020 from the grid layers on future threats to biodiversity. We required simple selection rules that, based on expert knowledge, give a reasonable range of future transformation rates across the CFR. Given that, with the exception of some areas affected by urbanization, it is unlikely that a single threat will completely transform all the natural areas in a grid cell by 2020, we chose the following approach: For our high future habitat transformation scenario, we regarded an area as transformed in 2020 if it is threatened by at least two of the three different threats at any levels. For our low future habitat transformation scenario, we regarded an area as transformed in 2020 if it is threatened by at least two of the three different future threats at high levels (i.e. highly threatened by urbanization, threatened by agriculture and highly threatened by alien plant invasions). Depending on the scenario, all areas within the existing reserves and proposed megareserves were however considered untransformed in 2020, irrespective of future threat levels. Finally, the future trans-

formed areas were combined with the current transformed areas to provide a single GIS layer of all areas transformed in 2020. We assumed that all currently transformed areas will still be transformed in 2020 (i.e. no restoration of currently transformed areas will take place).

At present 30.2% of the CFR is transformed and 13.5% is conserved (Rouget *et al.*, 2003b,c). By 2020, between 31.5% and 36.7% will be transformed, and between 13.5% and 29.6% will be conserved, depending on the scenario. The best- and worst-case future habitat transformation scenarios translate into mean transformation rates of 0.07% and 0.33% of the CFR per year, respectively or an additional 2.0% and 9.4% of the currently remaining natural areas in the CFR being transformed by 2020.

Red List assessment excluding and including future threats

We computed the Red List status for each taxon excluding and including future threats using a dBase IV programme (Borland Visual dBase Version 5.5a). We applied only criteria A and B of the new IUCN Red List Categories and Criteria, Version 3.1, which assess population reduction and geographic range parameters (Table 1) (IUCN, 2001, 2003). Criteria C, D and E were not applied because population size parameters and population viability analyses were not available for future scenarios. It should be noted that our simplistic, computer-based Red Listing approach is not comparable with a complete species-by-species IUCN Red List assessment.

Criterion A was applied using the proportion of transformed areas in a taxon's extent of occurrence (EOO) and area of occupancy (AOO) as a surrogate for population reduction (IUCN, 2001, 2003). The population reduction must be measured over the longer of 10 years or three generations (IUCN, 2001, 2003). The generation time of the Proteaceae is largely determined by the current fire cycle of about 10–30 years in most fynbos areas (Van Wilgen *et al.*, 1992). For the assessment excluding future threats, we set the reference period for population reduction to 60 years (1940–2000), which corresponds to a period of extensive habitat transformation in the CFR (Hoffman, 1997). For the

Fig. 3 (a) Currently transformed areas and additionally transformed areas for the best- and worst-case future habitat transformation scenario. (b) Increase in number of threatened Proteaceae taxa of the Cape Floristic Region per 1 min grid cell (2.9 km²) if worst-case land use change as predicted for 2020 is included in the Red List assessment. Boundaries of existing reserves are shown in gray. (c) Increase in number of threatened Proteaceae taxa of the Cape Floristic Region per 1 min grid cell (2.9 km²) if worst-case land use and climate change as predicted for 2020 is included in the Red List assessment. Boundaries of existing reserves and proposed megareserves are shown in gray.

assessment including future threats, we had to extend the reference period to 80 years (1940–2020), because detailed spatiotemporal information on the history of

habitat transformation is lacking. Criterion B was applied using a threshold of 20% transformed areas within a taxon's geographic range as an indicator of

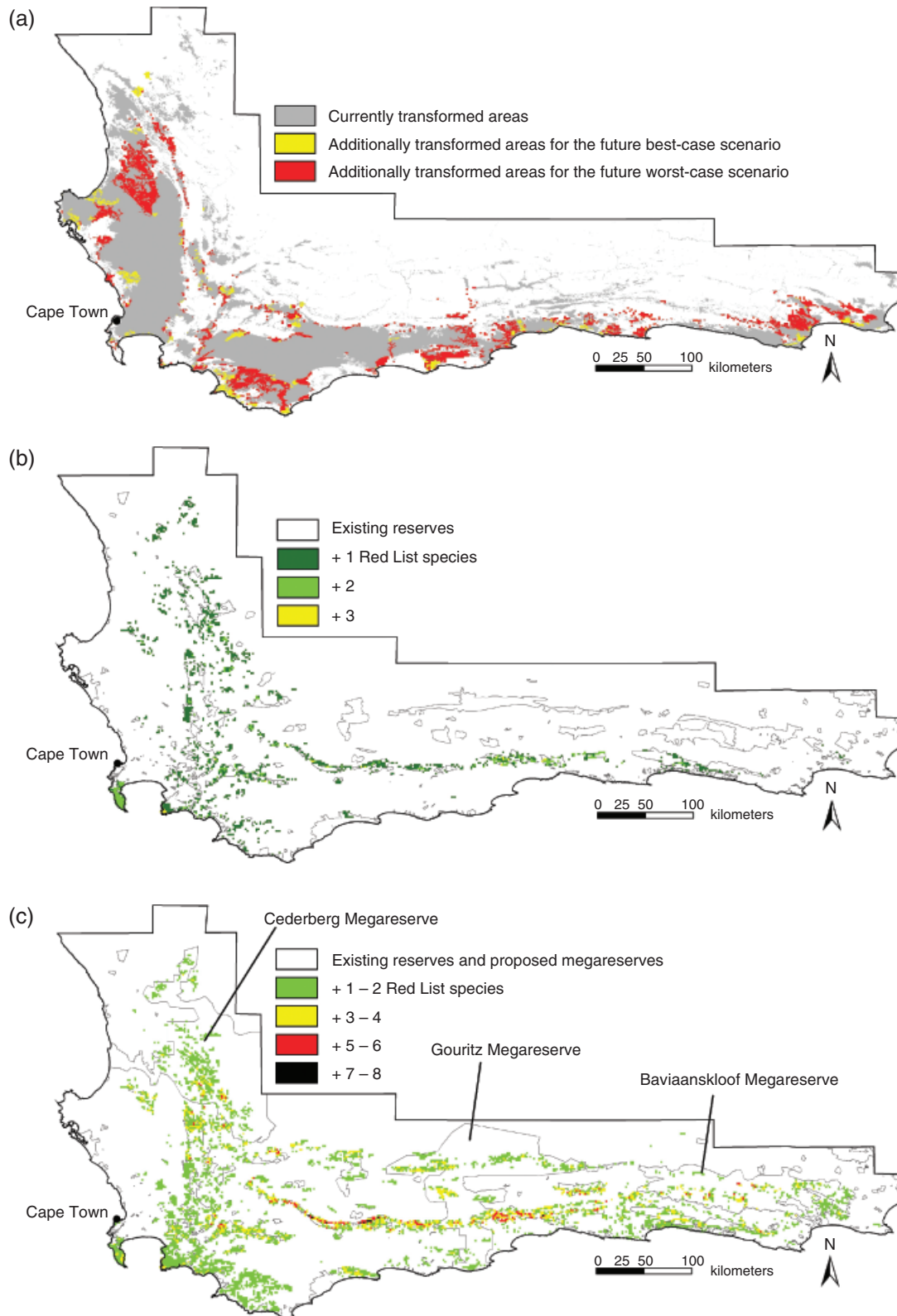


Table 1 Overview of the Red List criteria used to classify 227 Proteaceae taxa of the Cape Floristic Region into the Red List categories Critically Endangered (CR), Endangered (EN) and Vulnerable (VU) excluding and/or including future threats in the Red List assessment

	CR	EN	VU
<i>(A) Population reduction</i>			
Excluding future threats:			
A2. Population reduction estimated for the past based on (c) a decline in AOO or EOO estimated from ca. 1940 to 2000	> 80% transformed	> 50% transformed	> 30% transformed
Including future threats:			
A3. Population reduction projected for the future based on (c) a decline in AOO or EOO projected from 2000 to 2020	> 80% range loss	> 50% range loss	> 30% range loss
A4. Population reduction estimated for the past and projected for the future based on (c) a decline in AOO or EOO estimated and projected from ca. 1940 to 2020	> 80% transformed	> 50% transformed	> 30% transformed
<i>(B) Geographic range</i>			
Excluding and including future threats:			
B1. Either EOO or	< 100 km ²	< 5000 km ²	< 20000 km ²
B2. AOO	< 10 km ²	< 500 km ²	< 2000 km ²
and 2 of the following 3:			
(a) Number of locations (cells)	= 1	≤ 5	≤ 10
(b ii) Continuing decline in (i) EOO or (ii) AOO	> 20% transformed	> 20% transformed	> 20% transformed
(c iv) Extreme fluctuations in number of mature individuals	Assumed for reseeders but not for resprouters		

AOO is area of occupancy and EOO is extent of occurrence (IUCN, 2001, 2003).

continuing population decline. This threshold was based on expert knowledge from the Protea Atlas Project. We also assumed that all reseedling (in contrast to resprouting) study taxa experience extreme fire-related population fluctuations (Rebelo, 2001). Climate change impacts on species distributions were considered in the application of criteria A and B under the climate change scenarios.

Our program requires the following input data for each taxon: total area and percentage of transformed areas of EOO and AOO, percentage of range loss between now and 2020 under the climate change scenarios, number of locations (here grid cells occupied) and whether the taxon is a reseedling or resprouter. The data were compiled and computed using ArcGIS software (ESRI ArcGIS Version 8.2) and Arc Macro Language (AML) scripts. First, all GIS layers were projected to an Albers projection with World Geodetic System 1984 datum. Second, we generated species-

specific layers of the EOO by drawing the smallest possible circumscribing polygon around the species localities for current and future species distributions (IUCN, 2001), which were beforehand buffered by 500 m to adjust for potential locality pseudoreplicates arising from the strategy of temporal resampling (Pressey *et al.*, 2003). Third, we generated species-specific layers of the AOO by selecting only those cells of a 1 min grid containing species localities (IUCN, 2001). Clearly, AOO calculations are scale-dependent (IUCN, 2001; Hartley & Kunin, 2003; IUCN, 2003), and IUCN (2003) recommends a cell size of 4 km² for AOO calculations. Here, a 1 min grid with a cell size of about 2.9 km² was used because it was considered to be more appropriate for the often restricted and scattered ranges of the Proteaceae and determined by the similar resolution of the land use and climate change data. Finally, the number of locations, total area and percentage of transformed areas of EOO and AOO

were calculated for each taxon for now and 2020. Our program provides the following output data for each taxon: Red List categories and criteria met.

Taxa are classified as Critically Endangered (CR), Endangered (EN) or Vulnerable (VU) if they meet the relevant criteria shown in Table 1. These taxa are collectively referred to as Red List taxa or threatened taxa. Taxa that disappear from all their current localities under the climate change scenarios (i.e. they are expected to eventually become extinct) are listed as Extinct (EX). All remaining taxa are classified as Lower Risk (LR). Uplisting means a taxon is classified as more threatened than before, downlisting means a taxon is classified as less threatened than before.

Results

Red List status excluding and including future threats: general trends

For this study, all 227 Proteaceae taxa endemic to the CFR for which there were species-specific climate change predictions available were selected. If future threats are excluded from the Red List assessment there are 101 LR, 37 VU, 86 EN and three CR taxa among them (Table 2). According to our simulations, the current Red List status of the majority of study taxa does not change if future threats as predicted for 2020 are included in the assessment (Fig. 4a). However, between 3.5% and 29.1% of the 227 study taxa are

uplisted under the future scenarios. Land use change (scenarios –CC1 to –CC4) triggers uplistings for up to 25 taxa, but climate change (scenarios + CC1 to + CC4) triggers uplistings of up to three threat categories for up to 66 taxa (Table 2). Whereas 55.5% of the study taxa are classified as threatened (CR, EN or VU) if future threats are excluded from the assessment, 57.3% and 71.8% are listed as threatened or EX under the future best-(–CC1) and worst-case (+ CC4) scenarios, respectively (Fig. 4b).

With increasing severity of the future scenarios (i.e. towards high habitat transformation, low habitat conservation and including climate change impacts) (Fig. 2), the number of LR taxa decreases from 101 to 64 if future threats are included (Table 2). In contrast, the number of EX and CR taxa increases, particularly under the climate change scenarios, to four and 16, respectively, while the number of EN taxa rises from 86 to 101 under the worst-case scenario (+ CC4). Generally, high future habitat conservation (i.e. expansion of the existing reserve network by the three proposed mega-reserves) has almost no effect on the Red List status of the study taxa (only *Serruria fasciflora* is once listed as VU instead of EN with high conservation instead of low conservation), whereas high future habitat transformation has a major effect on the number of EN taxa. Climate change is, however, responsible for the majority of EX and CR listings.

Surprisingly, four taxa are downlisted under at least one of the climate change scenarios (Fig. 4a): *Leucadendron muirii*, *Protea acuminata*, *Protea recondita* and

Table 2 Overview showing for 227 modelled Proteaceae taxa of the Cape Floristic Region: Number of taxa uplisted (2–3 cat.: only those uplisted by two or three Red List categories), number of taxa per Red List category and Red List criteria met for Red List taxa, if future threats as predicted for 2020 are either excluded (present) or included in the Red List assessment (scenarios increase in severity from top to bottom: –CC is without climate change; + CC is with climate change)

	Taxa	Uplisted taxa	Uplisted taxa (2–3 cat.)	Number of taxa per Red List category					Red List criteria met for Red List taxa				Average number of criteria met
				LR	VU	EN	CR	EX	A3	A2/4	B1	B2	
Present	227	–	–	101	37	86	3	0	0	61	87	89	1.88
–CC1	227	8	1	97	37	88	5	0	0	66	85	86	1.82
–CC2	227	8	1	97	37	88	5	0	0	66	85	86	1.82
–CC3	227	25	8	88	32	101	6	0	0	83	86	89	1.86
–CC4	227	25	8	88	32	101	6	0	0	86	86	89	1.88
+ CC1	227	54	15	71	49	88	15	4	63	54	75	80	1.79
+ CC2	227	54	15	71	49	88	15	4	63	54	75	80	1.79
+ CC3	227	65	21	64	43	100	16	4	60	67	79	87	1.84
+ CC4	227	66	21	64	42	101	16	4	59	70	79	86	1.85

The different future scenarios are listed in Fig. 2. The Red List categories are Extinct (EX), Critically Endangered (CR), Endangered (EN), Vulnerable (VU) and Lower Risk (LR). The Red List criteria are listed in Table 1. More than one criterion was frequently met for an individual taxon justifying the assigned category: the last column shows the average number of Red List criteria met per Red List taxon.

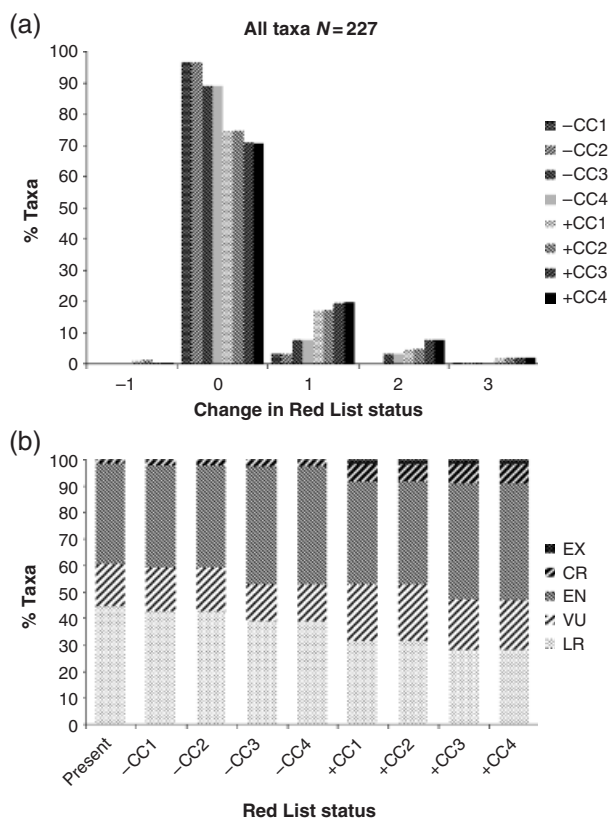


Fig. 4 (a) Change (number of categories downlisted or uplisted) in Red List status from a Red List assessment excluding future threats to one including future threats as predicted for 2020. (b) Red List status for 227 modelled Proteaceae taxa of the Cape Floristic Region if future threats as predicted for 2020 are either excluded (present) or included in the Red List assessment (scenarios increase in severity from left to right: -CC is without climate change; +CC is with climate change). The different future scenarios are listed in Fig. 2.

Serruria effusa. These taxa are downlisted because their smaller future ranges are relatively less transformed than their current ones while the reductions in range size are too small to affect their Red List status.

Taxa with highest threat status

Of the 20 taxa listed as CR or EX under any of the future scenarios (Table 2), climate change-induced range losses (+CC1 to +CC4) cause the critical endangerment of 11 taxa and the extinction of four taxa, namely *Leucadendron brunioides* var. *flumenlupinum*, *Leucadendron thymifolium*, *Leucospermum arenarium* and *Protea longifolia* var. *minor* (Table 3). In addition, low habitat transformation (-CC1 and -CC2) triggers the uplisting of two taxa into the CR category, with high habitat transformation (-CC3 and -CC4) adding another. The

remaining two taxa are already listed as CR if future threats are excluded and do not change their Red List status under any of the future scenarios. There is a geographic bias in the 20 taxa with 70% occurring on the West Coast and in the Southwest, whereas only 30% are from the South Central and Northwest.

Taxa with largest changes in threat status

In total, there are 22 different taxa that are uplisted by two or three Red List categories under at least one of the future scenarios (Fig. 4a). It is noteworthy that all threats can individually result in such uplistings. For example, *Diastella thymelaeoides* subsp. *meridiana* is a taxon from the Southwest that is uplisted from LR to CR under the best-case scenario (-CC1) and all other future scenarios (Table 3). On the other hand, the Red List status of taxa such as *Leucadendron 'touwsvriviensis'* (undescribed species) and *Leucospermum spathulatum* is only affected under the climate change scenarios (+CC1 to +CC4) (Table 3). Overall, climate change causes 16 of the highest uplistings, four times acting in concert with high habitat transformation, while high and low habitat transformation alone trigger five and one of the highest uplistings, respectively. Of the 22 different taxa, 82% are from the South Central and Southwest, and only 9% each from the West Coast and Northwest.

Which Red List criteria determined the Red List status of threatened taxa?

On average, nearly two (1.84) of the four Red List criteria applied here concurrently determined a threatened taxon's Red List status (Table 2). The total number of criteria met rises with increasing severity of the future scenarios. The geographic range criteria B1 and B2 were met the most, but with increasing severity of the future scenarios the population reduction criteria A3 and A4 were met increasingly. Criterion A3 indicates climate change impacts, whereas criterion A4 indicates land use change impacts (Table 1). Comparing the relative importance of these future threats based on Table 2 is difficult as both land use and climate change affect criterion B.

How are taxa from different regions affected?

Overall, the Proteaceae in the Southwest and, in particular, on the West Coast are more threatened than those in the South Central and Northwest (Fig. 5a). On the West Coast, the number of EX and CR taxa rises from 9.5% if future threats are excluded to 38.1% under the worst-case scenario (+CC4), and at least 90.5% of the taxa are classified as threatened or EX in any case.

Table 3 All taxa that are listed as Extinct (EX) or Critically Endangered (CR) of 227 modelled Proteaceae taxa of the Cape Floristic Region if future threats as predicted for 2020 are either excluded (present) or included in the Red List assessment (scenarios increase in severity from left to right: –CC is without climate change; +CC is with climate change)

Taxon	Present	Future scenarios				Cause	Region
		–CC1	–CC3	+CC1	+CC3		
		–CC2	–CC4	+CC2	+CC4		
<i>Diastella proteoides</i>	EN	EN	EN	CR	CR	CC	West Coast
<i>Diastella thymelaoides</i> subsp. <i>meridiana</i>	LR	CR	CR	CR	CR	LT	Southwest
<i>Leucadendron argenteum</i>	EN	CR	CR	CR	CR	LT	Southwest
<i>Leucadendron brunioides</i> var. <i>flumenlupinum</i>	EN	EN	EN	EX	EX	CC	West Coast
<i>Leucadendron levisanus</i>	EN	EN	CR	EN	CR	HT	West Coast
<i>Leucadendron remotum</i>	EN	EN	EN	CR	CR	CC	Northwest
<i>Leucadendron stellare</i>	CR	CR	CR	CR	CR	–	West Coast
<i>Leucadendron stelligerum</i>	EN	EN	EN	CR	CR	CC	Southwest
<i>Leucadendron thymifolium</i>	CR	CR	CR	EX	EX	CC	West Coast
<i>Leucadendron</i> ‘ <i>touwsrivierenses</i> ’ (undescribed sp.)	LR	LR	LR	CR	CR	CC	South Central
<i>Leucospermum arenarium</i>	EN	EN	EN	EX	EX	CC	West Coast
<i>Leucospermum conocarpodendron</i> subsp. <i>conocarpodendron</i>	CR	CR	CR	CR	CR	–	Southwest
<i>Leucospermum muirii</i>	EN	EN	EN	CR	CR	CC	South Central
<i>Leucospermum parile</i>	EN	EN	EN	CR	CR	CC	West Coast
<i>Leucospermum patersonii</i>	EN	EN	EN	CR	CR	CC	Southwest
<i>Leucospermum spatulatum</i>	LR	LR	LR	CR	CR	CC	Northwest
<i>Paranomus abrotanifolius</i>	EN	EN	EN	CR	CR	CC	South Central
<i>Protea convexa</i>	LR	LR	LR	CR	CR	CC	South Central
<i>Protea longifolia</i> var. <i>minor</i>	EN	EN	EN	EX	EX	CC	Southwest
<i>Serruria linearis</i>	EN	EN	EN	CR	CR	CC	West Coast

The different future scenarios are listed in Fig. 2. The other Red List categories are Endangered (EN), Vulnerable (VU) and Lower Risk (LR). The causes of uplistings are climate change (CC), high habitat transformation (HT) and low habitat transformation (LT).

In the Southwest, up to 92.5% of the taxa are either threatened or EX under the worst-case scenario (+CC4), but a few taxa remain in the LR category, and only 6.5% are classified as EX or CR under the climate change scenarios (+CC1 to +CC4). In contrast, 51.1% and 49.2% of the taxa in the Northwest and South Central remain in the LR category even under the worst-case scenario (+CC4). Moreover, no extinctions occur in these regions, but 4.4% of the taxa in the Northwest and 6.6% of the taxa in the South Central are listed as CR under the climate change scenarios.

How are taxa from different altitudes affected?

Overall, high-altitude taxa are less threatened than low-altitude taxa (Fig. 5b). Under the worst-case scenario (+CC4) for 2020, 94.9% of the taxa that occur below a mean altitude of 400 m at present are either threatened or EX, followed by 77.8% of the taxa between 400–800 m. In contrast, 43.6% of the taxa between 800–1200 m and 71.1% of the taxa above 1200 m remain in the LR category. Moreover, all extinctions and 14 out of

16 listings under the CR category involve taxa from below 400 m. Under the climate change scenarios (+CC1 to +CC4), the largest increases in the proportion of threatened vs. not-threatened taxa occur however above 400 m in the areas with the least threatened taxa to date. Under the worst-case scenario (+CC4), the proportion of threatened taxa increases roughly sixfold above 1200 m and twofold between 800 and 1200 m, whereas below 400 m it rises only from 88.5% to 94.9%.

Mapping the distribution of today's vs. tomorrow's threatened taxa

Comparing the mapped distribution of today's vs. tomorrow's threatened taxa (i.e. taxa classified as threatened excluding and including future threats) allows us to identify areas in which the number of these species of concern increases because of future land use and climate change as projected here. We therefore subtracted the number of threatened taxa per grid cell computed by a Red List assessment excluding

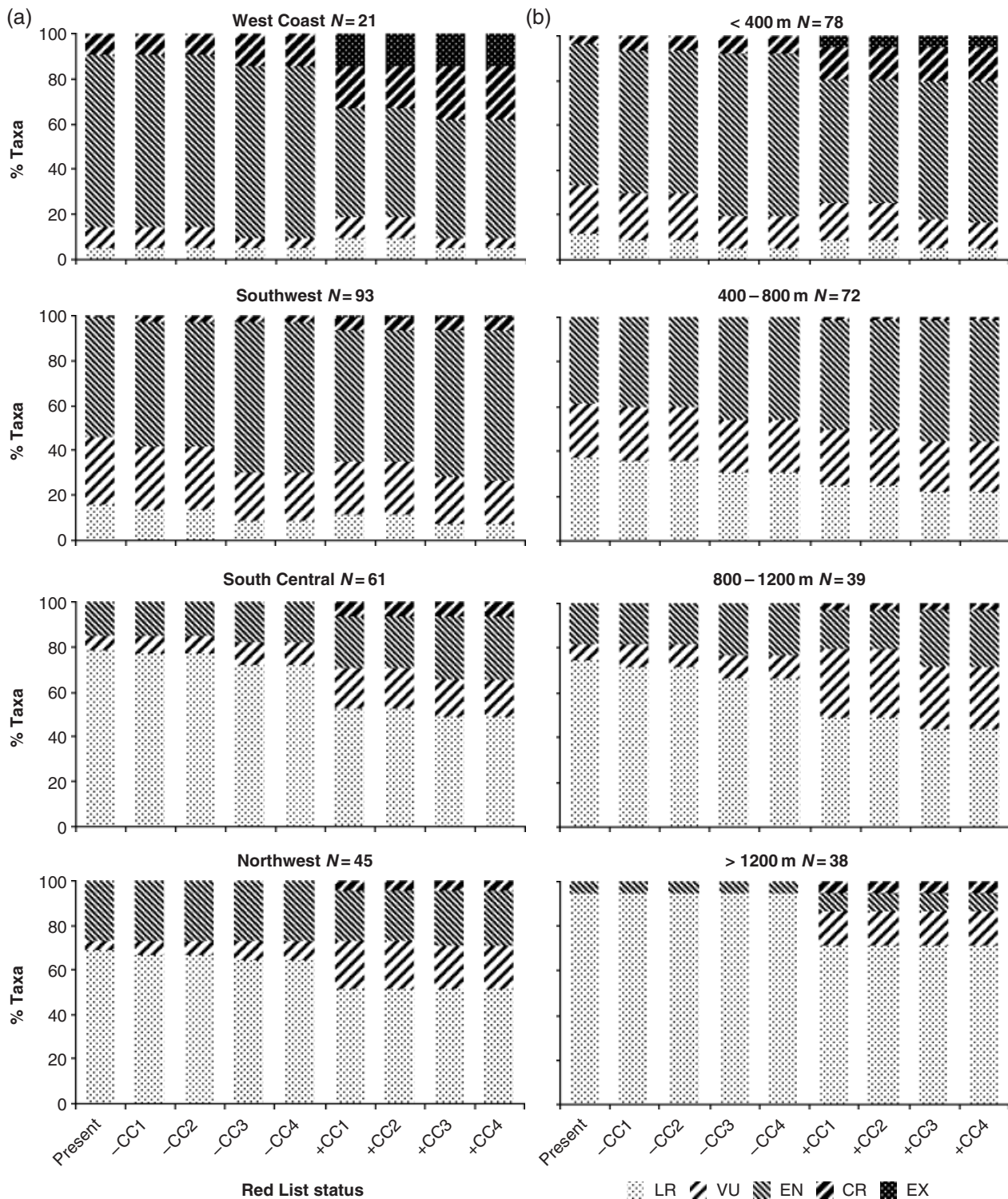


Fig. 5 Red List status for Proteaceae taxa from different regions (a) and altitudes (b) of the Cape Floristic Region if future threats as predicted for 2020 are either excluded (present) or included in the Red List assessment (scenarios increase in severity from left to right: –CC is without climate change; +CC is with climate change). The different future scenarios are listed in Fig. 2.

future threats from the number of threatened or EX taxa per grid cell under the future worst-case scenarios with (+CC4) and without (–CC4) climate change (Fig. 3b, c). Our maps show that, without climate change, the number of threatened taxa increases by up to three taxa per grid cell, with the Cape Peninsula, Pringle Bay area (southeast of Cape Town) and some inland mountain areas most affected by high habitat transformation. In contrast, with climate change, the number of threatened taxa increases by up to eight taxa per grid cell. Considerable changes occur especially in inland mountain areas, and to a lesser extent in coastal mountain areas such as the Cape Peninsula.

Discussion

For conservation planning and action, it is important to be able to assess how, where and when future threats could affect species persistence. Assessing this should facilitate the required shift from reactive to proactive conservation approaches. Our study estimates how many Proteaceae in the CFR could be affected (as shown by a change in their Red List status) by future threats that are usually not included in current IUCN Red List assessments. This was achieved by incorporating environmental conditions predicted by land use and climate change models for 2020 into a Red List assessment. The timeframe of this study allows us to evaluate our predictions in the near- to medium-term. A particular strength of our approach is the integration of both land use and climate change as major drivers of species endangerment and extinction. Furthermore, we integrated predicted species distributions (i.e. rather hypothetical potential ranges) with actual species distributions to assess future climate change impacts in the real world. Despite its limitations, our study serves as a first estimation of the 'shadow extinction risk' from future land use and climate change not evident in current IUCN Red Lists. The expected changes are substantial, affecting up to 29% of the 227 taxa, and they differ in their spatial and causal components. Species and regions not considered as threatened if future threats are excluded appear to be severely affected by land use and climate change in the future. Our results could be characteristic for many species and regions in and beyond the CFR (Burgman, 2002). Finally, our approach can in a sense be seen as a test of the value of planned interventions for conservation.

General limitations of our study

Our study should be seen as a first estimation of what could happen to the Red List status of the Proteaceae in the CFR by 2020. We regard our results as giving a

reasonable range of future trends based on our current knowledge of the potential impacts of future land use and climate change on the threat status of the Proteaceae at a regional scale. Our approach utilizes data that introduce uncertainty and, together with our assumptions, models and scenarios, should be interpreted accordingly. Caution should be used if our results are used to infer impacts for other taxa or at other temporal or spatial scales. In the next section we investigate possible reasons for underestimation or overestimation of the impacts of future threats on the Proteaceae.

Reasons for underestimation of the impacts of future threats

We did not assess some of the highly threatened range-restricted Proteaceae taxa in the CFR because it was impossible to correctly model their distributions and, thus, there were no species-specific climate change predictions available for them. Hence, our 227 study taxa include less than 10% of the taxa considered to be CR according to the Protea Atlas Project, which contains records for 377 Proteaceae taxa. All CR taxa are, however, expected to be still highly threatened or more so in 2020, and thus our study may seriously underestimate the overall impacts of future threats. Our study does not include the impacts of future threats other than land use and climate change, but commercial or casual plant harvesting is a serious threat to some of the charismatic Proteaceae (Rebello, 1992; Cowling *et al.*, 1997). Some of the existing reserves may not persist, and habitat transformation may not be prevented in both the existing reserves and proposed megareserves, as assumed here. For instance, invasive alien plants might further invade existing conservation areas, some of which are already densely invaded (Rouget *et al.*, 2003b). Possible positive feedbacks between land use and climate change (Root *et al.*, 2003; Opdam & Wascher, 2004) are not considered here. Climate change could however increase the fire frequency and intensity, putting Proteaceae that take up to 15 years to flower for the first time at risk, or change the fire season (Van Wilgen *et al.*, 1992). Climate change may also facilitate alien invasions with implications for indigenous species (Dukes & Mooney, 1999; McCarty, 2001). The predicted increase in invasive alien grasses in the CFR, for instance, could radically accelerate the fire cycle and, in turn, the extinction risk of the Proteaceae (Richardson *et al.*, 2000; Brooks *et al.*, 2004).

Reasons for overestimation of the impacts of future threats

We assessed only Proteaceae taxa endemic to the CFR, but more widespread taxa are expected to be less

threatened according to the Red List criteria. However, over 95% of Proteaceae are endemic to the CFR (Goldblatt & Manning, 2002). Most importantly, some plants could persist outside their potential future ranges in the near- to medium-term, because most Proteaceae are relatively long lived and some have long-lived seed banks (Cowling *et al.*, 1997; Rebelo, 2001), resulting in time lags between climate change and species responses. Some Proteaceae may also be more flexible with regard to their bioclimatic and edaphic constraints than assumed in the models (Davis & Shaw, 2001; Pearson & Dawson, 2003; Hampe, 2004), and could persist permanently outside their potential future ranges. With increasing atmospheric CO₂ concentrations, some Proteaceae may thrive, although the nutrient-poor environments of the CFR are likely to limit plant growth even in a CO₂-rich world (Stock & Midgley, 1995). Some Proteaceae may, both with or without human facilitation, successfully migrate and colonize new areas of their potential future ranges, thereby violating our assumption of no migration. The potential future ranges of about 16% of the taxa are in fact larger than their potential ranges at present. Potential ranges are however limited predictors of species distributions in the real world (Pearson & Dawson, 2003) – all Proteaceae occupy far fewer grid cells at present than predicted by their potential ranges, and an actual range expansion requires successful migration and colonization. However, many Proteaceae will have problems with ‘keeping up with climate change’ because of their limited dispersal ability compared with the rate at which the climate is predicted to change (Schurr *et al.*, 2004). Habitat fragmentation is a further obstacle to migration in the CFR (Rouget *et al.*, 2003c), and invasive alien plants, which are better dispersers, are likely to increasingly outcompete the indigenous Proteaceae (Richardson *et al.*, 2000). Habitat transformation in species ranges may however not necessarily result in population declines, as assumed here, and indigenous Proteaceae could persist without notable population declines in areas only moderately invaded by invasive alien plants (Latimer *et al.*, 2004). Except for the proposed megareserves no other future conservation actions were considered here. The existing reserve network in the CFR is, however, likely to be continuously enlarged and enhanced not only by the proposed megareserves, increasingly meeting the needs of threatened species (Cowling *et al.*, 2003; Rouget *et al.*, 2003b). Finally, lessons drawn from our study and others should lead to appropriate conservation action that may prevent some of the potential impacts of land use and climate change predicted here.

Implications of our study for the IUCN Red List Categories and Criteria

Comparing the Red List status of the Proteaceae excluding and including future threats highlights how many more threatened species could potentially be found in the CFR if we consider future threats to biodiversity. We suggest that the consideration of future threats is underutilized in current IUCN Red List assessments (Burgman, 2002), although the Red List criteria explicitly allow for this (IUCN, 2001, 2003). For proactive conservation planning and action, this information could be vital, but a lack of data or uncertainty about future threats may be reasons for not considering them. The introduction of a separate Red List reporting criterion for species potentially affected by future threats, particularly climate change, could be a solution to this problem. Our methodology needs to be further developed, but it could provide a first protocol for how to assess potential impacts of future threats.

We generally consider the IUCN Red List Categories and Criteria sensitive enough to detect changes in the threat status of species. It has been noticed, however, that changes in the threat status of species are often caused by changes in the knowledge of taxonomic classification or geographic distribution rather than actual changes in extinction risk (Burgman, 2002; Possingham *et al.*, 2002; Donaldson, 2004), which is not the case in our study. Both the interdependent nature and clear-cut thresholds of some Red List criteria, and our simplistic, computer-based Red Listing approach could however mask more subtle changes than the ones reported here. It is not easy to identify whether this would be because of flaws in our Red Listing approach or in the Red List criteria. The application of criterion A, which requires information on population reduction over specific timeframes (IUCN, 2001, 2003), poses probably the biggest problem, as detailed spatiotemporal information on the history of habitat transformation is lacking. This is probably the case in many parts of the world. In the absence of alternatives, it is also difficult to determine whether or not the proportion of transformed areas in species ranges is an appropriate surrogate for population reduction under the criteria A and B.

The variability and uncertainty of most species data make consistent Red List assessments already difficult (Akçakaya *et al.*, 2000). Incorporating future threats into Red List assessments adds further uncertainty, which is linked to the land use and climate change predictions, for instance. Nonetheless, it has been suggested that an index of threat be added to the Red List criteria, for example human population density (Harcourt & Parks,

2003). According to McKee *et al.* (2003), the number of threatened mammal and bird species in an average nation is expected to increase by 7% by 2020, as predicted by human population growth alone. According to our study, the proportion of threatened Proteaceae taxa could rise on average by 9.3% (range 1.8–16.3%) in the CFR by 2020. These numbers cannot be compared, because they are derived from different indicators of future threats, for instance. It is also important to note that the crude number of threatened vs. not-threatened species can mask significant changes within the threat categories. For South Africa, McKee *et al.* (2003) predict, in fact, downlistings of mammals and birds because of a projected national human population decline, but admit that continuing land use and climate change cast doubt on this prediction. On a local scale, human population will still continue to grow in areas such as Cape Town, where many threatened plants occur (Rebelo, 1992; Richardson *et al.*, 1996).

Both approaches highlight the desire to foresee future changes in the threat status of species (Manne & Pimm, 2001). Our study should inspire both the Red Listing and modelling community to further explore such approaches. For risk assessments of other taxa and regions, it would be vital to identify basic criteria of future endangerment and extinction, particularly as it is unlikely that species-specific models such as ours will ever be available for all taxa and regions. Our study highlights however that future threats – particularly when acting in concert – affect even Proteaceae with common characteristics so differently, that it is difficult to generalize our results based on species distributions, for example. Comparable studies, addressing this question in a more formal framework, are required not only for Proteaceae but also other taxa and regions to allow for generalizations. Finally, anticipating future threats to biodiversity should not only improve current risk assessments, but also our ability to take appropriate conservation action now, and thus the new Red List for southern African Proteaceae, which is currently being compiled, takes into account our results.

Applications of our study in conservation planning and action

This study is currently our best estimation of how the Red List status of the Proteaceae in the CFR could change in the near- to medium-term. Thus, it should be considered cautiously in conservation planning and action. The Proteaceae have been widely used for conservation planning in the CFR (Rebelo & Siegfried, 1992; Cowling *et al.*, 2003). Recently, retention targets were formulated for the Proteaceae, taking into account

the future threats from agriculture, alien plant invasions and urbanization (Pressey *et al.*, 2003). It would also be desirable to integrate spatially explicit information on climate change impacts such as range shifts and losses into conservation planning and reserve site-selection (Araújo *et al.*, 2004; Pyke *et al.*, 2005; Williams *et al.*, 2005). Range shifts and losses could considerably change the distributions of the Proteaceae and, in turn, their sensitivity to the impacts of 'conventional' threats (Opdam & Wascher, 2004; Hannah *et al.*, 2005).

Conservation planning and action requires the prioritization of species and regions of concern for reserve site-selection, scheduling and monitoring conservation efforts (Margules & Pressey, 2000). Information on where the number of threatened species increases because of future land use and climate change could be integrated into conservation planning as an additional GIS layer. Monitoring should be focused on Proteaceae that are expected to face the highest threat status or the largest changes in threat status, including the four taxa that are predicted to become extinct. Some of these taxa are not considered as species of concern according to current Red List assessments, which usually do not include future threats, but this will change with the new Red List for southern African Proteaceae. Low-altitude regions along the coast should be prioritized for conservation because they host more of the potentially most threatened species than high-altitude regions in the interior. Required migration distances (i.e. the distances that taxa have to migrate to reach their potential future ranges in a changing climate) are longer in flat, coastal areas than in mountain areas, where steeper climatic gradients prevail. This is particularly worrying as low-lying coastal areas already host the vast majority of threatened Proteaceae (Rebelo, 1992; Richardson *et al.*, 1996). Information such as that provided here on species and regions of concern are useful for the establishment of early warning systems. Threatened species profiling (i.e. the identification of common biological and ecological characteristics of threatened species) may further improve our understanding of combinations of traits and threats that predispose species to endangerment and extinction (Trinder-Smith *et al.*, 1996; Purvis *et al.*, 2000; Manne & Pimm, 2001).

When it comes to conservation action, the question is: where, and when, to start with which species? Our results may not only assist in locating and scheduling actions through prioritization, but also in detecting the most appropriate actions for certain species and regions, because they indicate the likely causes of future endangerment and extinction. For example, *Diastella thymelaoides* subsp. *meridiana* is uplisted from LR to CR even with low habitat transformation but is

not further affected by high habitat transformation and climate change (Table 3). Therefore, systematic habitat conservation in its current geographic range appears to be an appropriate action for this taxon. On the other hand, *Leucadendron brunioides* var. *flumenlupinum* is only affected by climate change, being uplisted from EN to EX under the climate change scenarios (Table 3), because at present it occurs outside of its potential future range. For this taxon, translocation from its current localities to its potential future range could be considered as an appropriate action to avoid its future extinction (Rutherford *et al.*, 1999). In practice, implications for the biodiversity in the target area have to be examined first, however (Rutherford *et al.*, 1999), to avoid adverse effects such as the hybridization of sister taxa, in this case with *Leucadendron brunioides* var. *brunioides*. This study should be used to pilot intervention strategies for Proteaceae predicted to be affected by climate change. Lessons drawn from monitoring and intervening, through translocation, species predicted here to be most affected, should be used to determine whether modelling future climate change impacts is useful for conservation.

Surprisingly, the three proposed megareserves ameliorate the threat status of only one taxon, which severely questions their importance for the conservation of Proteaceae. The megareserves coincide only marginally with the current distributions of the Proteaceae, and most of their newly conserved areas will be in the lowlands of the interior, whereas in these regions most of the Proteaceae are concentrated in the uplands. Therefore, other measures than the successful implementation of these megareserves are required for the Proteaceae. This does however not discredit the importance of the megareserve approach for other biodiversity patterns and processes (Cowling *et al.*, 1999; Younge, 2000; Rouget *et al.*, 2003a), some of which have not been considered at all in previous conservation planning, in contrast to the Proteaceae. Upland Proteaceae are less threatened at present because the mountain areas are less transformed and more conserved (Rouget *et al.*, 2003b,c). Lowland Proteaceae, in contrast, occur largely along the coasts and around Cape Town, where high human population densities, more fragmented and transformed areas, and less conservation areas are concentrated (Richardson *et al.*, 1996). Isolated reserves without latitudinal or altitudinal gradients and corridors, or without cautious translocation strategies, will not be a sustainable solution for the conservation of Proteaceae severely affected by climate change (Halpin, 1997; Hannah *et al.*, 2005). Given that climate change appears to cause significant changes in the Red List status of the Proteaceae, climate change-integrated conservation

strategies involving local mitigation and adaptation measures are as much needed as global actions such as the reduction of greenhouse gas emissions (Hannah *et al.*, 2002a,b).

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