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The diversity of biotic interactions complements functional and phylogenetic facets of biodiversity

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- 17 Summary

Taxonomic, functional and phylogenetic diversities are important facets of biodiversity. Studying them together has improved our understanding of community dynamics, ecosystem functioning and conservation values ¹⁻³. In contrast to species, traits, and phylogenies, the diversity of biotic interactions has so far been largely ignored as a biodiversity facet in large-scale studies. This neglect represents a crucial shortfall because biotic interactions shape community dynamics, drive important aspects of ecosystem functioning 4-⁷, provide services to humans, and have intrinsic conservation value ^{8,9}. Hence the diversity of interactions can provide crucial and unique information with respect to other diversity facets. Here, we leveraged large datasets of trophic interactions, functional traits, phylogenies and spatial distributions of >1000 terrestrial vertebrate species across Europe at a 10km resolution. We computed the diversity of interactions (Interaction Diversity, ID) in addition to functional (FD) and phylogenetic diversities (PD). After controlling for species richness, surplus and deficits of ID were neither correlated with FD nor with PD, thus representing unique and complementary information to the commonly studied facets of diversity. A threedimensional mapping allowed for simultaneously visualizing different combinations of ID-FD-PD. Interestingly, the spatial distribution of these diversity combinations closely matched the boundaries between ten European biogeographic regions, and revealed new, interaction-rich areas in the European Boreal region and interaction-poor areas in central Europe. Our study demonstrates that the diversity of interactions adds new and ecologically relevant information to multi-facetted, large-scale diversity studies with implications for understanding eco-evolutionary processes and informing conservation planning.

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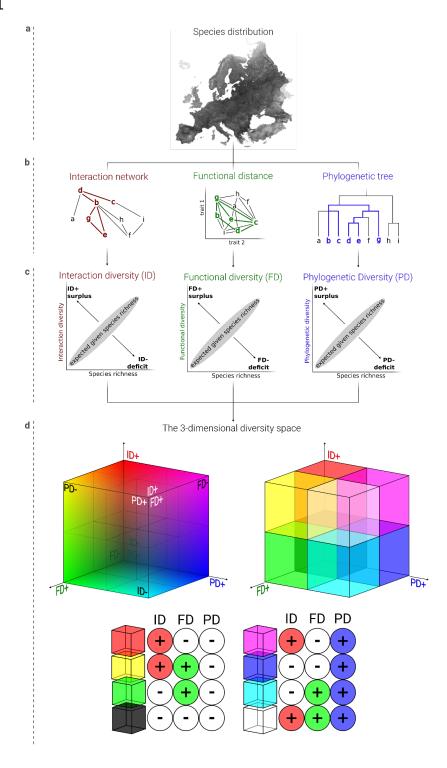


Figure 1. Conceptual workflow for a joint analysis of phylogenetic, functional, and interaction diversity. (a) Occurrences and probability of presence for 1149 terrestrial vertebrate species on 117,000 10×10km cells across Europe are combined with (b) the phylogenetic tree, a set of functional traits, and the trophic interactions of species. (c) We combined species distribution with phylogenetic, functional and trophic species attributes to compute local terrestrial vertebrate diversities using Hills numbers (q = 0, i.e "richness") and statistically corrected the diversity values by the local species richness. Note that the expected relationships (gray ellipses) are not necessarily linear. (d) We projected the diversity values in a 3-dimensional space with each axis representing a diversity facet and a color in the Red-Blue-Green space (x = PD / blue, y = FD / green, z = ID / red), and discretized particular types of combinations based on surplus and deficits of each diversity. Red identifies surpluses of ID and FD and PD (ID>0, FD<0), PD<0); Yellow identifies surpluses of ID and FD associated with deficits in PD (ID>0, FD>0, PD<0); Green identifies deficits in ID and PD associated with FD surpluses (ID<0, FD>0, PD<0); Black identifies deficits in ID, PD and FD, Pink identifies surpluses of PD associated with ID and FD deficits; Light blue identifies surpluses of PD and FD associated with ID deficits; white identifies surpluses in ID, PD and FD.

Biodiversity -the diversity of life on Earth- was originally used to refer to species diversity, but it is now used to reflect a multi-faceted concept ³. Given the evidence that species diversity alone cannot appropriately describe community assembly, ecosystem functioning and variation in community composition ¹⁰, several complementary measures of biodiversity have emerged in the last three decades ^{1,2}. The most important are the diversity of species' evolutionary histories (i.e phylogenetic diversity, PD) and their ecological functions (i.e functional diversity, FD), but while PD and FD are becoming central to many studies ^{11,12}, the diversity of biotic interactions (i.e interaction diversity, ID) has been poorly considered as a biodiversity facet in large-scale studies (but see refs ^{8,13}). This is a major gap since biological interactions are tightly linked to species coexistence ¹⁴, ecosystem productivity and functioning ^{4–7}.

In its simplest form, ID is the total number of interactions shared by all species of a given assemblage⁹. Interactions considered can be of different types and nature, e.g antagonistic (competition for resources), mutualistic (pollination ^{15,16}), or trophic (predation ¹⁷). Although the concept of interaction diversity is not novel ^{8,13} and has its own methodological tools ^{18,19}, the lack of information available on biotic interactions ^{20,21} has limited its study across large taxonomical and spatial scales ^{22–26}. Here, we leveraged unique and valuable data combining spatial distributions²⁷ (Figure 1.a), trophic interactions (Figure 1.b), functional traits²⁸, and phylogenies²⁹ of most terrestrial vertebrate species in Europe ³⁰ at a 10 km resolution. Within each 10km cell, we computed interaction diversity (ID, as the number of trophic interactions), functional diversity (FD, as the sum of functional pairwise Gower distances between species in the cell), and phylogenetic diversity (PD, as the sum of the branch lengths of the phylogenetic tree containing all species present in the cell ¹) using Hill numbers ^{31 18}. We statistically corrected each diversity by the local species richness ^{32,33} in order to measure and map deficits and surpluses of ID, FD, and PD (Figure 1.c). We also investigated the correlation and complementarity between the three facets, and created a 3-dimensional diversity space that reveals different local combinations of ID-FD-PD (Figure 1.d) and their distribution across biogeographical regions in Europe.

Surpluses and deficits of diversities

Trophic networks of terrestrial vertebrates found within 10km cells in Europe contained up to 4834 trophic interactions with an average of 1958 interactions across cells (Figure 2a). Once corrected for species richness, ID ranged from a deficit of -942 interactions (1667 observed interactions with 202 species involved) to a surplus of +968 interactions (3730 interactions with 210 species involved, see Supplemental information - Trophic network examples). Because highly connected assemblages are often considered as the signature of functional and resilient ecosystems 34,35, areas with high ID are important from a conservation point of view 34,36. Further, comparing spatial distributions of surplus and deficit IDs with those of FD or PD can complement our understanding of community dynamics and underlying processes. Because phylogenetic and trait data contain information about evolutionary history and species niches, the spatial distribution of their diversity (Figure 2b-c) is thought to hold the signature of the eco-evolutionary drivers that shape biodiversity patterns ^{37–39}. For example, for a given species richness, observed surplus of FD (Figure 1d, and Figure 2b, green color) could result from competitive exclusion between species with similar traits, while a deficit of FD might result from environmental filtering constraining the range of locally viable traits or hierarchical competition where a given set of traits is the best adapted locally 40. PD surplus (Figure 1d and Figure 2c, dark blue color) could result from slow extinction rates of old and distant lineages (i.e., museums of biodiversity 41), and PD deficit from rapid recent speciation (i.e., cradles of biodiversity). ID surplus and deficit brings additional information, as observed ID surplus (Figure 1d and Figure 2a, red color) indicates particularly dense or long trophic networks, such as those emerging from high levels of omnivory and intraguild predation 42, or from bottom-up control when large amounts of basal resources sustain longer trophic chains and the presence of top predators. ID deficits can result from weakened topdown control when top predators are absent from local assemblages, for example following human-induced removal 43-45.

Overall, the different facets of diversity are shaped by eco-evolutionary drivers which are not mutually exclusive ⁴⁶. Any combination of ID-FD-PD could potentially exist locally and bring complementary information to the others, although one can expect the facets of diversity to be (partly) correlated when similar drivers influence multiple diversity facets. We showed that FD and PD were clearly and positively

correlated (Figure 3b). This correlation is due to the fact that species tend to retain their ancestral traits through evolution ^{38,47–49}, and suggests an important effect of evolution and phylogenetic niche conservatism on biodiversity patterns ^{47,50,51}. While one could expect ID to be related to FD (because of the link between trait similarity and competition for resources) or PD (because biotic interactions can drive the (co)evolutionary history of the species ^{46,52}), this was not what we observed (Figure 3c-d). Instead, ID represented unique and complementary information to the commonly studied facets of diversity.

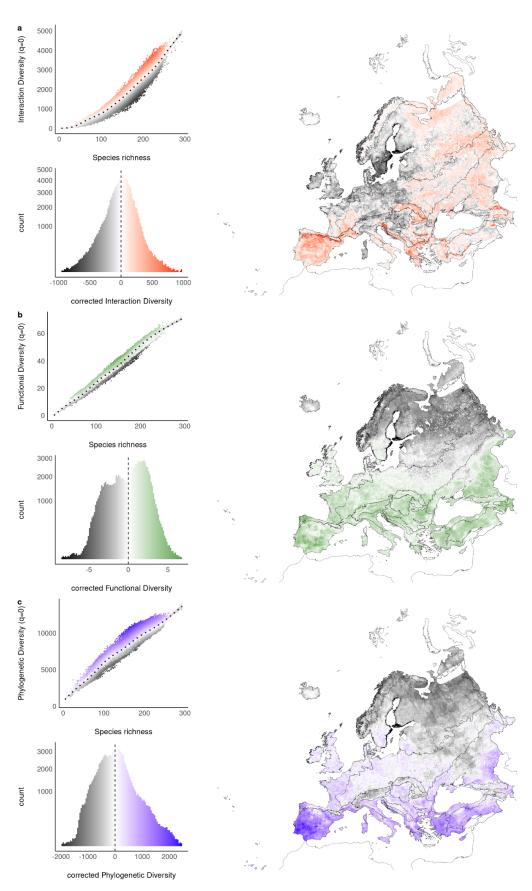


Figure 2. Patterns of Interaction diversity ID (a, in red), Functional diversity FD (b, in green), and Phylogenetic diversity PD (c, in blue). Top left: Relationship between each diversity facet and the species richness. Dotted lines show relationships as fitted by Generalized Additive Models. Bottom left: Distribution of deficits and surpluses of diversities, where model residuals correspond to "corrected diversity" values with deficits (dark shades) and surpluses (red for ID, green for FD, blue for PD). Right: spatial distribution of corrected values for each biodiversity facet, color corresponds to distributions on the left.

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Distribution of diversity combinations

To investigate the congruence between the interaction, functional and phylogenetic facets of biodiversity, we created a 3-dimensional space where each dimension represents one diversity facet. We further attributed a color channel for each diversity facet (red = ID, green = FD, blue = PD) to visualize all possible combinations of biodiversity facets (Figure 1d). Each combination of three color channels (Red, Blue, Green) resulted in a particular color in the RGB color space that corresponds to a given combination of three diversity facets, and allowed us to identify a continuum of ID-FD-PD combinations (Figure 1d). We also interpreted particular types of combinations by discretizing colors based on the combinations of surplus and deficits of each diversity facet (Figure 1d).

This joint analysis of diversity facets highlighted various local combinations of ID-FD-PD, with all kinds of combinations being observed in different proportions (Figure 3a). The most commonly observed combinations were ID surpluses with FD and PD deficits (covering 21.8% of the total study area); surpluses in ID, FD, and PD (white, 21.6%); surpluses of FD and PD with deficits in ID (light blue 19.6%); and deficits in ID, FD, and PD (black, 17.3%), which is consistent with the positive correlation observed between FD and PD (Figure 3b). The spatial structure of diversity combinations aligned well with many boundaries of European biogeographical regions (Figure 3a), a striking spatial congruency considering that the identification and delimitation of bioregions are based on geographic distribution of vegetation types 53. Beyond species distribution, biodiversity facets such as phylogenetic diversity already have been shown to match some ecological regions across the globe 54. ID strongly varies between different regions (e.g. between the Mediterranean region and the Alps, or between the Continental region and the Carpathian mountains) and thus further refines boundaries between them. These results suggest that species interactions (along with species co-occurrences and phylogeny) could have a strong structuring effect on (bio)regional species pools. Such a question, however, would require a deeper analysis based on the turnover of interactions within and between regions as regional diversity is connected to local diversity by the turnover in composition between locations. Interestingly, the mapping of diversity combinations also revealed the specificity of several sub-regions within their biogeographical region, e.g the Balkan peninsula sub-region in the Mediterranean region, or the Carpathian mountains in the Alpine region. These results further highlight that biotic interaction diversity adds new and independent information and that a dense network of trophic interactions can occur in areas of poor functional and phylogenetic diversity.

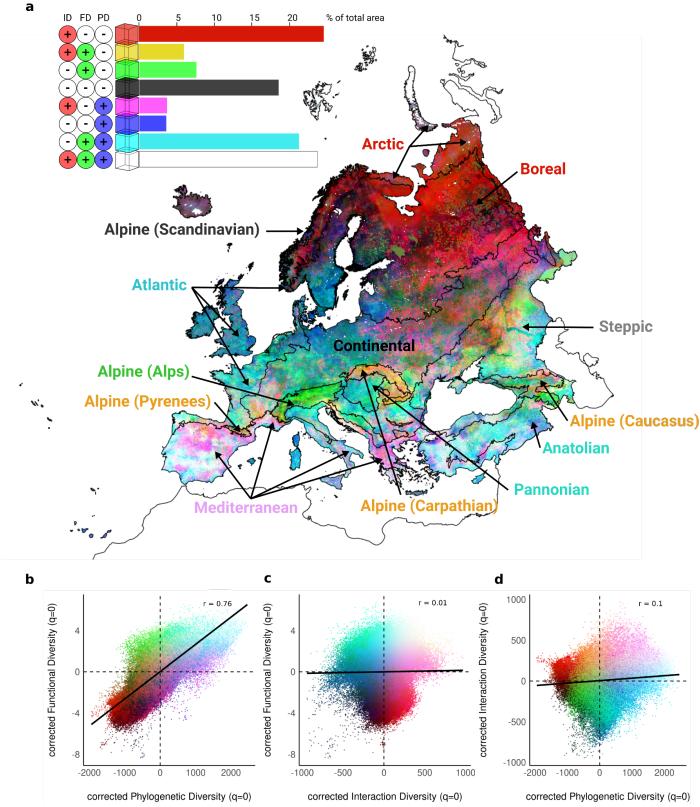


Figure 3. (a) Spatial projection of the 3-dimensional diversity space. In the top left barplot, we created 8 discrete categories based on the combinations of deficits (-) and surpluses (+) of each diversity and reported the number of cells falling in each category. In the map, points are colored by their location in the Red-Green-Blue 3-dimensional color space, with each diversity facet corresponding to a distinct channel: Red channel = Interaction Diversity, Green Channel = Functional Diversity, Blue channel = Phylogenetic Diversity. black shows lowest ID-FD-PD values, white shows highest ID-FD-PD, and so on for each combination. Black lines show the boundaries of the European biogeographical regions. (b) Pair plot of corrected FD (y-axis) VS corrected FD (y-axis) VS corrected ID (y-axis) VS corrected ID (y-axis) VS corrected PD (x-axis), (c) corrected PD (x-axis), respectively.

Southern Europe showed strong diversity surpluses in all diversity facets (white / light color shades in Figure 3a), which confirms the Mediterranean bioregion as a multifaceted biodiversity hotspot ^{55,56}. This result shows that, for a given number of species, local assemblages of Mediterranean terrestrial vertebrate species were particularly rich in terms of ecological strategies, contained long evolutionary history, and had

particularly dense trophic networks. In the Mediterranean basin, the warm climate and the geographical proximity with Africa and Asia explains the high diversity of amphibians and reptiles, as well as the presence of unique evolutionary lineages, leading to high functional and phylogenetic diversities compared to the rest of Europe (Supplemental information - Groups and species distributions across Europe). In addition to these high levels of functional and phylogenetic diversities, the Mediterranean region showed surpluses in interaction diversity, in particular in the subregion of the Balkan peninsula. The densely connected trophic networks observed in the Mediterranean region resulted from (i) numerous top predators in this region (Supplemental information - Maps of relevant network properties) previously identified as birds, felids, and snakes predating upon small reptiles and rodents ²²; and (ii) to a lesser extent from a high degree of omnivory in the Iberian peninsula ²³.

Conversely, the northernmost areas tended to show low levels of diversities (black areas in North of Scandinavia and Iceland, Figure 3a). The Boreal and Arctic bioregions showed deficits in functional and phylogenetic diversities, but tended to sustain surpluses in interaction diversity (red areas, Figure 3a). In these regions, FD deficits were likely to be driven by the cold climate constraining the range of functional traits that can be found in these regions, and similarly for PD via trait conservatism. In particular, the fact that cold temperature limits the presence of ectotherms (amphibians and reptiles) in high latitudes reduces functional and phylogenetic diversities, in line with the expected effect of environmental filtering on these diversity facets ^{37,38}. The consideration of ID brings additional and complementary information since FD-PD deficits are associated with ID surpluses in Northern Europe. The presence of ID surpluses in the Boreal and Arctic bioregions likely resulted from a high degree of omnivory (Supplemental information - Maps of relevant network properties), which is known to increase trophic network connectance ^{26,57}. Species that live under high latitudes tend to be trophic generalists ²³ because the higher seasonality in high latitudes promotes the evolution of larger niche breadth, in accordance with the latitude–niche breadth hypothesis ^{57,58}

Within the Alpine bioregion, different mountain ranges displayed contrasting diversity combinations. The marked differentiation between the Alps and the Carpathian mountains subregions is a striking example supporting the consideration of interaction diversity in biodiversity studies and conservation biogeography. These two mountain ranges located in Central Europe are part of the same Alpine bioregion, which partly explains their similarity in terms of functional surpluses and phylogenetic deficits (Figure 2b-c). Based on functional and phylogenetic diversities alone, these two mountain ranges would be considered as similarly diverse - but they are markedly different in terms of interaction diversity. The Carpathians displayed a clear ID surplus in (Figure 3, yellow), while the Alps were deficitary (Figure 3, green). The proximate cause of such difference was the rarity of top predators in the Alps compared to the Carpathians (see <u>Supplemental information - Maps of relevant network properties</u>). Human influence likely explains this discrepancy because many apex-predators (bears, wolves, lynx) that are often trophic generalists are still present in the Carpathians, while they were exterminated in the Alps ⁵⁹.

Potential drivers of diversity facets

While environmental filtering is likely to drive the decrease of FD and PD observed in high latitudes, ID might be more influenced by human activities than climate. As such, local deficits of trophic interactions appeared as a marker of high human impact across Europe. This is in line with the negative correlation between connectance and human influence previously reported for the same study system ²³, and suggests that the diversity of interactions is influenced by different drivers than functional and phylogenetic diversity. It is, however, noteworthy that other studies reported higher connectance in more human impacted systems ^{9,46}. Indeed, the human-induced relative increase of generalist intermediate predators could counterbalance the decrease in ID due to the loss of a few top predators. The human influence on large-scale diversity has been considered and studied in terms of phylogeny and traits ⁶⁰. However, its consequences on large-scale patterns of interaction diversity have been largely overlooked, although they are probably stronger. Indeed, human activities have been (and still are) particularly detrimental to large-bodied species ⁶⁰⁻⁶³. While this

observation is generally viewed as a trait-induced consequence (humans are more detrimental to larger animals), it might also be a trophic-induced consequence (humans are more detrimental to apex and generalist predators) ^{45,64}.

The importance of interaction diversity

A clear understanding of the impact of human activities on ID has yet to emerge. More generally, ID is likely to be highly context and taxa dependent, and the understanding of its multi-scale drivers represents a research agenda for the years to come. Among others, the Eltonian shortfall is one big challenge that currently limits the description of ID in many parts of the world where information on biotic interactions is lacking ²¹. Here, we overcame this challenge for trophic interactions by inferring local interactions from species distributions and their known potential trophic interactions from the literature and expert knowledge (as commonly done, see for example refs ^{19,25}). While this approach overestimates interactions at a given time, "realized" and "potential" number of interactions are very likely to converge in the long term. On the contrary, a field sampling approach would underestimate the realized ID. This underestimation can be quite severe and a massive sampling effort is required to detect most interactions ⁶⁵. Combining both approaches (inferring interactions from a metanetwork and species distribution, vs. observing interactions), and comparing their accuracy across a range of temporal and spatial scales will provide valuable insights in community ecology and biogeography ⁶⁶.

Although ID patterns appear robust to data depletion and spatial contexts (see <u>Supplemental information</u>-Robustness of diversity patterns), whether the patterns described in this study can be extrapolated to other biomes remains an open question. For example, our conclusions from European terrestrial vertebrates might not hold true for tropical rainforests which shelter many trophic specialist species with narrow ecological niches (but comprehensive data on traits and interactions are lacking). Nonetheless, we argue that interaction diversity is a particularly valuable facet for biogeography and conservation planning. Although this view has been empirically challenged ³⁶, more densely connected trophic networks are generally considered as desirable from a conservation point of view ⁹. Areas with surpluses of interactions represent interaction networks that are expected to be more robust to cascading species extinctions ³⁴, and consequently more resilient to perturbations. Coupled with its apparent sensitivity to human activities ²³, interaction diversity might be viewed as a marker of both ecosystem degradation and resistance to future degradation. We argue that a general consideration of interaction diversity as an important and meaningful diversity facet alongside the functional and phylogenetic diversities should be a priority for macroecology and conservation biogeography.

STAR Methods

Study area and data

 Study area. The study area, hereafter referred to as "Europe", included the entire European subcontinent (with Macaronesia and Iceland) plus Anatolia to include a complete picture of the North Mediterranean coast (Figure 1a). The study area was divided into 117,000 cells on a 10×10 km equal-size area grid (ETRS89). Within the study area, we considered ten biogeographical regions defined by the European Environment Agency ⁶⁷: Alpine, Anatolian, Arctic, Atlantic, Boreal, Black Sea, Continental, Macaronesia, Mediterranean, and Steppic. These bioregions are large scale ecological units based on an interpretation of geobotanical data ⁶⁸, and represent areas with homogeneous ecological context.

Species distributions. We extracted the distributions for all terrestrial vertebrates naturally occurring within the study area from Maiorano et al. (2013). Species distributions for 509 bird, 288 mammal, 250 reptile and 104 amphibian species were mapped by combining the IUCN extent of occurrence for each species with their habitat requirements. A species was considered potentially present in a 10×10 km cell if the grid cell met the three following criteria: i) is within the species extent of occurrence, ii) contains at least one 300x300m area of primary habitat for the species, i.e. habitat where the species can persist (defined by experts and published literature) and iii) meets species requirements in terms of elevation and distance from water. A full description of species distribution data and definition of primary habitat can be found in Maiorano et al. 2013 ²⁷. In addition, we used the percentage of primary habitat of the species in each cell as a proxy for the probability to find the species in a random locality within this cell. For example, we considered that if the primary habitat of a species covered 80% of the cell, the probability to find the species in a random locality of the cell was 0.8. As such, it represents a proxy for the probability of presence of the species within the cell and was used as a weight in the entropy-based diversity measures (i.e when q=1) provided as supplementary analyses.

Functional traits. We gathered biological trait data from Thuiller et al. 2015 28, excluding traits describing diet (and thus trophic interactions) and traits for larvae and juveniles. Our analysis was based on four lifehistory and ecological traits common to mammals, amphibians, birds and reptiles. The only quantitative trait was body mass [grams, log-transformed]. The three other traits were multichoice nominal variables coded by binary values. Feeding behavior was coded by four binary columns: opportunistic feeder, active hunter, browser, grazer. Nesting location was coded by eleven binary columns: tree/hole/fissure in bark, ground, building/artificial, underground water, cave/fissures/burrows, lodge. brooks/springs/small rivers, puddles/ponds/pools/small lakes, brackish waters. Activity time was coded by four binary columns: nocturnal, crepuscular, diurnal, arrhythmic. These traits were selected because they represent informative niche dimensions linked to the use and acquisition of resources in space and time, and are related to ecosystem functioning ^{69,70}. A thorough description of traits and the list of publications where the data were gathered is available in supplementary material from ²⁸ available at https://royalsocietypublishing.org/doi/suppl/10.1098/rstb.2014.0005.

We computed the pairwise dissimilarities (distances) of this trait matrix using a mixed variable coefficient of distance (using function dist.ktab in ade4) that generalizes Gower's general coefficient of distance to allow the treatment of various statistical types of variables when calculating distances ⁷¹. Euclidean distance was used for body mass, and Jaccard index was used for the four other multichoice nominal variables (S3 coefficient of in Gower and Legendre 1986 ⁷²).

Phylogenetic tree. We used the 100 phylogenetic trees for European terrestrial vertebrates assembled and published by Roquet at al. 2014 ²⁹. We chose these phylogenetic trees as they are the only species-level phylogenies encompassing all european vertebrates, and have already been valuably used to depict phylogenetic diversity in this context vertebrates in the past ^{28,32,73,74}.

Trophic networks. We used data on species trophic interactions from the metaweb of European terrestrial vertebrates, (Tetra-EU 1.0, ³⁰). This metaweb is based on expert knowledge, published information and field guides. Potential trophic links between a predator and a prey were identified from published accounts of their observation, morphological similarities between potential prey and literature-referenced prey or -in the absence of this information- the diet of the predator's sister species. The metaweb of European terrestrial vertebrates contained 1,164 species and a total of 50,408 potential trophic interactions. The full dataset and methods description can be found in ref ³⁰.

In order to maximize the species coverage for each diversity, we allowed for different sets of species to be used to compute ID, FD and PD. For ID we retained 1149 species for which we had information on their European distribution range and trophic interactions; for FD we retained 1009 species for which we had information on their European distribution range and functional traits; for PD we retained 993 species for which we had information on their European distribution range and phylogeny. This varying set of species should have low impact on the assessment of diversities as ID, FD and PD were corrected by their corresponding taxonomic richness to compute surpluses and deficits. In order to investigate the potential bias resulting from the variation of species coverage across space and diversities, we computed diversities based on the same set of 884 species for which we had all shared information. The resulting diversity patterns were similar when considering the 884 species or varying set of species (see Supplemental information - Diversities based on the same set of 884 species).

Diversity measures and the 3-dimensional diversity space

Within each 10x10km cell, we used Hill numbers ³¹ to compute FD, PD, and ID. In this framework, diversity values are converted into effective numbers of species, the Hill numbers. When considering taxonomic diversity, the effective number of species is the number of equally abundant species necessary to produce the observed value of diversity (an analogue to the concept of effective population size in genetics). This approach has then been generalized to incorporate species phylogenetic relatedness and species functional distances. We used the framework from Chao et al. 2014 ³¹ implemented in the R package hillR for phylogenetic and functional diversity and in the package econetwork ¹⁸ for interaction diversity. We computed each diversity as a Hill number analogous to a measure of richness by setting q=0 (ignoring abundance). The ID richness was the sum of trophic links formed by the species present in the cell, the FD richness was the sum of functional pairwise gower distances between species in the cell, and the PD richness was the mean sum of the branch lengths of the phylogenetic tree connecting all species present in the cell ¹ across the 100 trees.

We focused our study on richness-based results (q=0) as they are the easiest to interpret, but we also analyzed and showed results based on Shannon entropy in the <u>Supplemental information - Results based on Shannon entropy</u>. To compute the results as a Shannon entropy, we set q=1 and used the % of species' primary habitat within the cell as the probability to find the species in the cell. More precisely, when q=1 the ID entropy is the Shannon entropy over the interaction weights (product of the two species abundances), the FD is the Shannon entropy of effective number of species-pairs with unit-distance between species, and the PD is the mean Shannon entropy of the effective total branch length across the 100 trees. More details on the calculations of FD and PD can be found in Chao et al. 2014 ³¹, and in Ohlmann et al. 2019 ¹⁸ for ID.

We corrected FD, PD, ID richness and Shannon entropy for the number of species in the cell (i.e taxonomic richness) based on the set of species used to compute each diversity. We fitted a thin plate spline regression, a particular Generalized Additive Model (GAM), to predict each diversity measure from species richness. The residuals of each model (one for each diversity facet and order q) were retained as the species richness corrected value of the diversity, with positive residuals considered as surplus and negative residuals considered as deficits given the species richness ^{32,33}. In other words, a deficit (or surplus) indicates a lower (or higher, respectively) diversity value than expected given the local species richness (Figure 1).

To investigate the congruence between the interaction, functional and phylogenetic facets of biodiversity, we created a 3-dimensional space where each dimension represents one diversity facet. In order to visualize all possible combinations of biodiversity facets, we attributed a color channel for each diversity facet (red = ID, green = FD, blue = PD) where the residual values for each diversity were rescaled to 0-255 value in the corresponding color channel (Figure 1). Hence, each combination of three color channels (Red, Blue, Green) results in a particular color in the RGB color space that corresponds to a given combination of three diversity facets, and allows us to identify a continuum of ID-FD-PD combinations depicted in figure 1.d. We can also interpret particular types of combinations by discretizing colors based on the combinations of surplus and deficits of each diversity. As shown in Figure 1d and Figure 3a, Red identifies surpluses of ID and FD associated with deficits in FD and PD (ID>0, FD<0, PD<0); Yellow identifies surpluses of ID and FD associated with deficits in PD (ID>0, FD>0, PD<0); Green identifies deficits in ID and PD associated with FD surpluses (ID<0, FD>0, PD<0); Black identifies deficits in ID, PD and FD, Pink identifies surpluses of ID and PD associated with FD deficits; Dark blue identifies surpluses of PD associated with ID and FD deficits; Light blue identifies surpluses of PD and FD.

SUPPLEMENTAL INFORMATION

Supplemental Information includes six appendix and can be found at:

AUTHOR CONTRIBUTIONS

- P.G and W.T. conceived the study, with early advice from L.O.C, C.B, G.P. and T.M. P.G. performed all analyses and wrote the first version of the manuscript with inputs from W.T, L.O.C, C.B, G.P. and T.M. All authors contributed substantially to the interpretation of the results and to the writing of the manuscripts and its revisions.
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References

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- 404 1. Faith, D.P. (1992). Conservation evaluation and phylogenetic diversity. Biol. Conserv. 61, 1–10.
- 405 2. Tilman, D. (2001). Functional diversity. Encyclopedia of biodiversity 3, 109–120.
- 406 3. Pollock, L.J., O'Connor, L.M.J., Mokany, K., Rosauer, D.F., Talluto, M.V., and Thuiller, W. (2020).
 407 Protecting Biodiversity (in All Its Complexity): New Models and Methods. Trends Ecol. Evol. 35, 1119–
 408 1128.
- 409 4. Grime, J.P. (1997). Biodiversity and Ecosystem Function: The Debate Deepens. Science 277, 1260–410 1261.
- 5. Tilman, D., Isbell, F., and Cowles, J.M. (2014). Biodiversity and Ecosystem Functioning. Annu. Rev. Ecol. Evol. Syst. 45, 471–493.
- 413 6. Brose, U., Blanchard, J.L., Eklöf, A., Galiana, N., Hartvig, M., R. Hirt, M., Kalinkat, G., Nordström, M.C., O'Gorman, E.J., Rall, B.C., et al. (2017). Predicting the consequences of species loss using size-structured biodiversity approaches. Biol. Rev. Camb. Philos. Soc. *92*, 684–697.
- 416 7. Schneider, F.D., Brose, U., Rall, B.C., and Guill, C. (2016). Animal diversity and ecosystem functioning in dynamic food webs. Nat. Commun. *7*, 12718.
- Thompson, J.N. (1997). Conserving Interaction Biodiversity. In The Ecological Basis of Conservation: Heterogeneity, Ecosystems, and Biodiversity, S. T. A. Pickett, R. S. Ostfeld, M. Shachak, and G. E. Likens, eds. (Springer US), pp. 285–293.
- 421 9. Tylianakis, J.M., Laliberté, E., Nielsen, A., and Bascompte, J. (2010). Conservation of species interaction networks. Biol. Conserv. *143*, 2270–2279.
- 423 10. Hooper, D.U., Solan, M., Symstad, A., Diaz, S., Gessner, M.O., Buchmann, N., Degrange, V., Grime, P., Hulot, F., Mermillod-Blondin, F., et al. (2002). Species diversity, functional diversity and ecosystem functioning. Biodiversity and ecosystem functioning: synthesis and perspectives, 195–208.
- De Palma, A., Kuhlmann, M., Bugter, R., Ferrier, S., Hoskins, A.J., Potts, S.G., Roberts, S.P.M.,
 Schweiger, O., and Purvis, A. (2017). Dimensions of biodiversity loss: Spatial mismatch in land-use impacts on species, functional and phylogenetic diversity of European bees. Divers. Distrib. 23, 1435–1446.
- 430 12. Schweiger, A.K., Cavender-Bares, J., Townsend, P.A., Hobbie, S.E., Madritch, M.D., Wang, R., Tilman, D., and Gamon, J.A. (2018). Plant spectral diversity integrates functional and phylogenetic components of biodiversity and predicts ecosystem function. Nat Ecol Evol *2*, 976–982.
- 13. Dyer, L.A., Walla, T.R., Greeney, H.F., Stireman, J.O., III, and Hazen, R.F. (2010). Diversity of Interactions: A Metric for Studies of Biodiversity: Interaction Diversity. Biotropica *42*, 281–289.
- 435 14. Jacquet, C., Moritz, C., Morissette, L., Legagneux, P., Massol, F., Archambault, P., and Gravel, D. (2016). No complexity–stability relationship in empirical ecosystems. Nat. Commun. 7, 1–8.
- 437 15. Burkle, L.A., and Alarcón, R. (2011). The future of plant-pollinator diversity: understanding interaction networks across time, space, and global change. Am. J. Bot. *98*, 528–538.
- 439 16. Burkle, L.A., Myers, J.A., and Belote, R.T. (2016). The beta-diversity of species interactions:
- Untangling the drivers of geographic variation in plant-pollinator diversity and function across scales.
- 441 Am. J. Bot. 103, 118–128.
- 442 17. Sandom, C., Dalby, L., Fløjgaard, C., Kissling, W.D., Lenoir, J., Sandel, B., Trøjelsgaard, K., Ejrnaes, R., and Svenning, J.-C. (2013). Mammal predator and prey species richness are strongly linked at
- 444 macroscales. Ecology *94*, 1112–1122.

- 18. Ohlmann, M., Miele, V., Dray, S., Chalmandrier, L., O'Connor, L., and Thuiller, W. (2019). Diversity indices for ecological networks: a unifying framework using Hill numbers. Ecol. Lett. *22*.
- 447 19. Poisot, T., Canard, E., Mouillot, D., Mouquet, N., and Gravel, D. (2012). The dissimilarity of species interaction networks. Ecol. Lett. *15*, 1353–1361.
- 449 20. Hortal, J., de Bello, F., Diniz-Filho, J.A.F., Lewinsohn, T.M., Lobo, J.M., and Ladle, R.J. (2015). Seven Shortfalls that Beset Large-Scale Knowledge of Biodiversity. Annu. Rev. Ecol. Evol. Syst. *46*, 523–549.
- 451 21. Poisot, T., Bergeron, G., Cazelles, K., Dallas, T., Gravel, D., MacDonald, A., Mercier, B., Violet, C., 452 Vissault, S., and Chapman, D. (2021). Global knowledge gaps in species interaction networks data. J. 453 Biogeogr. 48, 1552–1563.
- 454 22. O'Connor, L.M.J., Pollock, L.J., Braga, J., Ficetola, G.F., Maiorano, L., Martinez-Almoyna, C.,
 455 Montemaggiori, A., Ohlmann, M., and Thuiller, W. (2020). Unveiling the food webs of tetrapods across
 456 Europe through the prism of the Eltonian niche. J. Biogeogr. *47*, 181–192.
- 457 23. Braga, J., Pollock, L.J., Barros, C., Galiana, N., Montoya, J.M., Gravel, D., Maiorano, L.,
 458 Montemaggiori, A., Ficetola, G.F., Dray, S., et al. (2019). Spatial analyses of multi-trophic terrestrial
 459 vertebrate assemblages in Europe. Glob. Ecol. Biogeogr. 28, 1636–1648.
- 24. Cumming, G.S., Bodin, Ö., Ernstson, H., and Elmqvist, T. (2010). Network analysis in conservation biogeography: Challenges and opportunities. Diversity and Distributions *16*, 414–425.
- 462 25. Gravel, D., Baiser, B., Dunne, J.A., Kopelke, J.P., Martinez, N.D., Nyman, T., Poisot, T., Stouffer, D.B., 463 Tylianakis, J.M., Wood, S.A., et al. (2019). Bringing Elton and Grinnell together: a quantitative 464 framework to represent the biogeography of ecological interaction networks. Ecography *42*, 401–415.
- 465 26. Baiser, B., Gravel, D., Cirtwill, A.R., Dunne, J.A., Fahimipour, A.K., Gilarranz, L.J., Grochow, J.A., Li, D., Martinez, N.D., McGrew, A., et al. (2019). Ecogeographical rules and the macroecology of food webs. Glob. Ecol. Biogeogr. *28*, 1204–1218.
- 468 27. Maiorano, L., Amori, G., Capula, M., Falcucci, A., Masi, M., Montemaggiori, A., Pottier, J., Psomas, A., Rondinini, C., Russo, D., et al. (2013). Threats from climate change to terrestrial vertebrate hotspots in Europe. PLoS One *8*, e74989.
- 471 28. Thuiller, W., Maiorano, L., Mazel, F., Guilhaumon, F., Ficetola, G.F., Lavergne, S., Renaud, J., Roquet, C., and Mouillot, D. (2015). Conserving the functional and phylogenetic trees of life of European tetrapods. Philos. Trans. R. Soc. Lond. B Biol. Sci. *370*, 20140005.
- 29. Roquet, C., Lavergne, S., and Thuiller, W. (2014). One tree to link them all: a phylogenetic dataset for the European tetrapoda. PLoS Curr. 6.
- 476 30. Maiorano, L., Montemaggiori, A., Ficetola, G.F., O'Connor, L., and Thuiller, W. (2020). TETRA-EU 1.0: A species-level trophic metaweb of European tetrapods. Glob. Ecol. Biogeogr. *29*, 1452–1457.
- 478 31. Chao, A., Chiu, C.-H., and Jost, L. (2014). Unifying Species Diversity, Phylogenetic Diversity, 479 Functional Diversity, and Related Similarity and Differentiation Measures Through Hill Numbers. Annu. 480 Rev. Ecol. Evol. Syst. 45, 297–324.
- 481 32. Zupan, L., Cabeza, M., Maiorano, L., Roquet, C., Devictor, V., Lavergne, S., Mouillot, D., Mouquet, N., Renaud, J., and Thuiller, W. (2014). Spatial mismatch of phylogenetic diversity across three vertebrate groups and protected areas in Europe. Diversity and Distributions *20*, 674–685.
- 484 33. Safi, K., Cianciaruso, M.V., Loyola, R.D., Brito, D., Armour-Marshall, K., and Diniz-Filho, J.A.F. (2011). 485 Understanding global patterns of mammalian functional and phylogenetic diversity. Philos. Trans. R. 486 Soc. Lond. B Biol. Sci. *366*, 2536–2544.
- 487 34. Gilbert, A.J. (2009). Connectance indicates the robustness of food webs when subjected to species loss. Ecol. Indic. 9, 72–80.
- 489 35. Thébault, E., and Fontaine, C. (2010). Stability of ecological communities and the architecture of

- 490 mutualistic and trophic networks. Science 329, 853–856.
- 491 36. Heleno, R., Devoto, M., and Pocock, M. (2012). Connectance of species interaction networks and conservation value: Is it any good to be well connected? Ecol. Indic. *14*, 7–10.
- 493 37. McGill, B.J., Enquist, B.J., Weiher, E., and Westoby, M. (2006). Rebuilding community ecology from functional traits. Trends Ecol. Evol. *21*, 178–185.
- 495 38. Webb, C.O., Ackerly, D.D., McPeek, M.A., and Donoghue, M.J. (2002). Phylogenies and Community 496 Ecology. Annu. Rev. Ecol. Syst. 33, 475–505.
- 497 39. Münkemüller, T., Gallien, L., Pollock, L.J., Barros, C., Carboni, M., Chalmandrier, L., Mazel, F.,
 498 Mokany, K., Roquet, C., Smyčka, J., et al. (2020). Dos and don'ts when inferring assembly rules from
 499 diversity patterns. Glob. Ecol. Biogeogr. *164*, S165.
- 40. Mayfield, M.M., and Levine, J.M. (2010). Opposing effects of competitive exclusion on the phylogenetic structure of communities. Ecol. Lett. *13*, 1085–1093.
- 502 41. Stebbins, G.L. (1974). Flowering Plants: Evolution Above the Species Level (Belknap Press of Harvard University Press).
- 504 42. Duffy, J.E., Cardinale, B.J., France, K.E., McIntyre, P.B., Thébault, E., and Loreau, M. (2007). The functional role of biodiversity in ecosystems: incorporating trophic complexity. Ecol. Lett. *10*, 522–538.
- 506 43. Estes, J. a., Terborgh, J., Brashares, J.S., Power, M.E., Berger, J., Bond, W.J., Carpenter, S.R., Essington, T.E., Holt, R.D., Jackson, J.B.C., et al. (2011). Trophic downgrading of planet Earth. Science 333, 301–306.
- 44. de Visser, S.N., Freymann, B.P., and Olff, H. (2011). The Serengeti food web: empirical quantification and analysis of topological changes under increasing human impact. J. Anim. Ecol. *80*, 484–494.
- 511 45. Pauly, D., Christensen, V., V., Dalsgaard, J., Froese, R., and Torres, F., Jr (1998). Fishing down marine food webs. Science *279*, 860–863.
- 513 46. Tylianakis, J.M., and Morris, R.J. (2017). Ecological Networks Across Environmental Gradients. Annu. Sev. Ecol. Evol. Syst. 48, 25–48.
- 515 47. Wiens, J.J., and Graham, C.H. (2005). Niche Conservatism: Integrating Evolution, Ecology, and Conservation Biology. Annu. Rev. Ecol. Evol. Syst. *36*, 519–539.
- 517 48. Peterson, A.T., Soberon, J., and Sanchez-Cordero, V., V. (1999). Conservatism of ecological niches in evolutionary time. Science *285*, 1265–1267.
- 49. Lavergne, S., Mouquet, N., Thuiller, W., and Ronce, O. (2010). Biodiversity and Climate Change:
 Integrating Evolutionary and Ecological Responses of Species and Communities. Annu. Rev. Ecol.
 Evol. Syst. 41, 321–350.
- 522 50. Wiens, J.J., Ackerly, D.D., Allen, A.P., Anacker, B.L., Buckley, L.B., Cornell, H.V., Damschen, E.I., 523 Jonathan Davies, T., Grytnes, J.-A., Harrison, S.P., et al. (2010). Niche conservatism as an emerging 524 principle in ecology and conservation biology. Ecol. Lett. *13*, 1310–1324.
- 525 51. Münkemüller, T., Boucher, F., Thuiller, W., and Lavergne, S. (2015). Common conceptual and methodological pitfalls in the analysis of phylogenetic niche conservatism. Funct. Ecol. *29*, 627–639.
- 527 52. Williams, R.J., and Martinez, N.D. (2000). Simple rules yield complex food webs. Nature *404*, 180– 183.
- 529 53. Noirfalise, A. (1987). Map of the Natural Vegetation: Of the Member Countries of the European Economic Community and the Council of Europe.
- 531 54. Holt, B.G., Lessard, J.-P., Borregaard, M.K., Fritz, S.A., Araújo, M.B., Dimitrov, D., Fabre, P.-H., Graham, C.H., Graves, G.R., Jønsson, K.A., et al. (2013). An update of Wallace's zoogeographic

- regions of the world. Science 339, 74–78.
- 534 55. Myers, N., Fonseca, G. a. B., Mittermeier, R. a., Fonseca, G. a. B., and Kent, J. (2000). Biodiversity hotspots for conservation priorities. Nature *403*, 853–858.
- 536 56. Mittermeier, R.A., Turner, W.R., Larsen, F.W., Brooks, T.M., and Gascon, C. (2011). Global 537 Biodiversity Conservation: The Critical Role of Hotspots. In Biodiversity Hotspots: Distribution and 538 Protection of Conservation Priority Areas, F. E. Zachos and J. C. Habel, eds. (Springer Berlin 539 Heidelberg), pp. 3–22.
- 540 57. Cirtwill, A.R., Stouffer, D.B., and Romanuk, T.N. (2015). Latitudinal gradients in biotic niche breadth vary across ecosystem types. Proc. Biol. Sci. *282*.
- 542 58. Vázquez, D.P., and Stevens, R.D. (2004). The latitudinal gradient in niche breadth: concepts and evidence. Am. Nat. *164*, E1–19.
- 544 59. Kaczensky, P., Chapron, G., Von Arx, M., Huber, D., Andrén, H., and Linnell, J. (2013). Status, 545 Management and Distribution of Large Carnivores - Bear, Lynx, Wolf & Wolverine - in Europe (Verlag 546 nicht ermittelbar).
- 60. Fritz, S.A., Bininda-Emonds, O.R.P., and Purvis, A. (2009). Geographical variation in predictors of mammalian extinction risk: big is bad, but only in the tropics. Ecol. Lett. *12*, 538–549.
- 549 61. Gaston, K.J., and Blackburn, T.M. (1995). Birds, body size and the threat of extinction. Philos. Trans. 550 R. Soc. Lond. B Biol. Sci. *347*, 205–212.
- 551 62. Cardillo, M., Mace, G.M., Jones, K.E., Bielby, J., Bininda-Emonds, O.R.P., Sechrest, W., Orme, C.D.L., 552 and Purvis, A. (2005). Multiple causes of high extinction risk in large mammal species. Science *309*, 553 1239–1241.
- 554 63. Böhm, M., Williams, R., Bramhall, H.R., McMillan, K.M., Davidson, A.D., Garcia, A., Bland, L.M., 555 Bielby, J., and Collen, B. (2016). Correlates of extinction risk in squamate reptiles: the relative 556 importance of biology, geography, threat and range size. Glob. Ecol. Biogeogr. *25*, 391–405.
- 557 64. Brose, U. (2010). Improving nature conservancy strategies by ecological network theory. Basic Appl. Ecol. *11*, 1–5.
- 559 65. Chacoff, N.P., Vázquez, D.P., Lomáscolo, S.B., Stevani, E.L., Dorado, J., and Padrón, B. (2012). 560 Evaluating sampling completeness in a desert plant-pollinator network. J. Anim. Ecol. *81*, 190–200.
- 561 66. Fortuna, M.A., Nagavci, A., Barbour, M.A., and Bascompte, J. (2020). Partner Fidelity and Asymmetric Specialization in Ecological Networks. Am. Nat. *196*, 382–389.
- 563 67. European Environmental Agency, EEA (2016). Biogeographical regions. 564 https://www.eea.europa.eu/data-and-maps/data/biogeographical-regions-europe-3.
- 68. Roekaerts, M. (2002). The Biogeographical Regions Map of Europe. Basic principles of its creation and overview of its development (European Environmental Agency, EEA).
- 567 69. Luck, G.W., Lavorel, S., McIntyre, S., and Lumb, K. (2012). Improving the application of vertebrate 568 trait-based frameworks to the study of ecosystem services. Journal of Animal Ecology *81*, 1065–1076.
- 569 70. Sekercioglu, C.H. (2006). Increasing awareness of avian ecological function. Trends Ecol. Evol. *21*, 464–471.
- 571 71. Pavoine, S., Vallet, J., Dufour, A.-B., Gachet, S., and Daniel, H. (2009). On the challenge of treating various types of variables: application for improving the measurement of functional diversity. Oikos 118, 391–402.
- 574 72. Gower, J.C., and Legendre, P. (1986). Metric and Euclidean properties of dissimilarity coefficients. J. Classification *3*, 5–48.

- 576 73. Pearman, P.B., Lavergne, S., Roquet, C., Wüest, R., Zimmermann, N.E., and Thuiller, W. (2014).
 577 Phylogenetic patterns of climatic, habitat and trophic niches in a European avian assemblage. Glob.
- 578 Ecol. Biogeogr. 23, 414–424.
- 579 74. Saladin, B., Thuiller, W., Graham, C.H., Lavergne, S., Maiorano, L., Salamin, N., and Zimmermann, N.E. (2019). Environment and evolutionary history shape phylogenetic turnover in European tetrapods.
 - Nat. Commun. 10, 249.

592

593

- 582 75. Cadotte, M.W., Cardinale, B.J., and Oakley, T.H. (2008). Evolutionary history and the effect of biodiversity on plant productivity. Proc. Natl. Acad. Sci. U. S. A. *105*, 17012–17017.
- 76. Cadotte, M.W., Jonathan Davies, T., Regetz, J., Kembel, S.W., Cleland, E., and Oakley, T.H. (2010). Phylogenetic diversity metrics for ecological communities: integrating species richness, abundance
- and evolutionary history. Ecol. Lett. *13*, 96–105.
- 587 77. Davies, T.J., and Buckley, L.B. (2011). Phylogenetic diversity as a window into the evolutionary and biogeographic histories of present-day richness gradients for mammals. Philos. Trans. R. Soc. Lond. B Biol. Sci. *366*, 2414–2425.
- 78. Pollock, L.J., Thuiller, W., and Jetz, W. (2017). Large conservation gains possible for global biodiversity facets. Nature *546*, 141–144.
 - 79. Cadotte, M.W., Carscadden, K., and Mirotchnick, N. (2011). Beyond species: functional diversity and the maintenance of ecological processes and services. J. Appl. Ecol. 48, 1079–1087.
- 594 80. Díaz, S., and Cabido, M. (2001). Vive la différence: plant functional diversity matters to ecosystem processes. Trends Ecol. Evol. *16*, 646–655.
- 596 81. Cadotte, M.W., Dinnage, R., and Tilman, D. (2012). Phylogenetic diversity promotes ecosystem stability. Ecology 93, S223–S233.
- 598 82. Brun, P., Zimmermann, N.E., Graham, C.H., Lavergne, S., Pellissier, L., Münkemüller, T., and Thuiller, W. (2019). The productivity-biodiversity relationship varies across diversity dimensions. Nat. Commun. 10, 5691.
- 601 83. de Bello, F., Lavorel, S., Hallett, L.M., Valencia, E., Garnier, E., Roscher, C., Conti, L., Galland, T.,
 602 Goberna, M., Májeková, M., et al. (2021). Functional trait effects on ecosystem stability: assembling
 603 the jigsaw puzzle. Trends Ecol. Evol. *36*, 822–836.
- 604 84. Loreau, M. (2014). Reconciling utilitarian and non-utilitarian approaches to biodiversity conservation. 605 Ethics Sci. Environ. Polit. *14*, 27–32.
- 85. Mazel, F., Wüest, R.O., Gueguen, M., Renaud, J., Ficetola, G.F., Lavergne, S., and Thuiller, W. (2017). The Geography of Ecological Niche Evolution in Mammals. Curr. Biol. *27*, 1369–1374.
- 608 86. Chesson, P., and Kuang, J.J. (2008). The interaction between predation and competition. Nature *456*, 609 235–238.
- 610 87. Chase, J.M., Abrams, P.A., Grover, J.P., Diehl, S., Chesson, P., Holt, R.D., Richards, S.A., Nisbet, R.M., and Case, T.J. (2002). The interaction between predation and competition: a review and synthesis. Ecol. Lett. *5*, 302–315.
- 88. Stouffer, D.B., Sales-Pardo, M., Sirer, M.I., and Bascompte, J. (2012). Evolutionary conservation of species' roles in food webs. Science *335*, 1489–1492.
- 89. Poisot, T., Mouquet, N., and Gravel, D. (2013). Trophic complementarity drives the biodiversityecosystem functioning relationship in food webs. Ecol. Lett. *16*, 853–861.
- 90. Vane-Wright, R.I., Humphries, C.J., and Williams, P.H. (1991). What to protect?—Systematics and the agony of choice. Biol. Conserv. *55*, 235–254.
- 91. Isaac, N.J.B., Turvey, S.T., Collen, B., Waterman, C., and Baillie, J.E.M. (2007). Mammals on the

620 EDGE: conservation priorities based on threat and phylogeny. PLoS One 2, e296.

626

627

632

- 92. Rodrigues, A.S.L., and Gaston, K.J. (2002). Maximising phylogenetic diversity in the selection of networks of conservation areas. Biol. Conserv. *105*, 103–111.
- 93. Devictor, V., Mouillot, D., Meynard, C., Jiguet, F., Thuiller, W., and Mouquet, N. (2010). Spatial mismatch and congruence between taxonomic, phylogenetic and functional diversity: the need for integrative conservation strategies in a changing world. Ecol. Lett. *13*, 1030–1040.
 - 94. Peterson, A.T., Soberón, J., Pearson, R.G., Anderson, R.P., Martínez-Meyer, E., Nakamura, M., and Bastos Araujo, M. (2011). Ecological niches and geographic distributions (Princeton University Press).
- 628 95. Díaz, S., Settele, J., Brondízio, E., Ngo, H., Guèze, M., Agard, J., Arneth, A., Balvanera, P., Brauman, K., Butchart, S., et al. (2020). Summary for policymakers of the global assessment report on biodiversity and ecosystem services of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services (IPBES).